

1 **Dual effect of Se(IV) and bentonite microbial community interactions on the**
2 **corrosion of copper and Se speciation: implication on repository safety assessment**

3

4 Marcos F. Martinez-Moreno^{1,*}, Cristina Povedano-Priego¹, Mar Morales-Hidalgo¹, Adam
5 D. Mumford², Guillermo Lazuen-Lopez¹, Elisabet Aranda³, Ramiro Vilchez-Vargas⁴, Pier
6 L. Solari⁵, Yon Ju-Nam², Fadwa Jroundi¹, Jesus J. Ojeda², and Mohamed L. Merroun¹.

7

8 ¹*Faculty of Sciences, Department of Microbiology, University of Granada, Granada,*
9 *Spain.*

10 ²*Department of Chemical Engineering, Faculty of Science and Engineering, Swansea*
11 *University, Swansea, United Kingdom*

12 ³*Institute of Water Research, Department of Microbiology, University of Granada,*
13 *Granada, Spain*

14 ⁴*Medical Department II, University Hospital, Ludwig-Maximilians-Universität, Munich,*
15 *Germany*

16 ⁵*MARS Beamline, Synchrotron SOLEIL, L' Orme des Merisiers, Départementale 128,*
17 *Saint-Aubin, France*

18

19

20 **Keywords**

21 *Deep Geological Repository, selenium speciation, bentonite, bacteria, copper*
22 *biocorrosion*

23

24 **Abstract**

25 The Deep Geological Repository design, the internationally safest option for the long-
26 term disposal of high-level radioactive waste (HLW), features metal canisters encased in
27 compacted bentonite clay and embedded deep within a host rock. Despite presenting a
28 hostile environment for microorganisms, DGRs scenarios with favorable microbial-
29 activity conditions must be considered for the safety assessment of this disposal. This
30 study investigated the impact of Se(IV), as a natural analogue of ⁷⁹Se present in the HLW,
31 in anoxic microcosms of bentonite slurry spiked with a bacterial consortium and amended
32 with lactate, acetate, and sulfate as electron donors/acceptor. The addition of the bacterial
33 consortium promoted the rate of Se(IV) reduction to Se(0), while the tyndallization (heat-
34 shock) of bentonite slowed this process. Se(IV) reduced the relative abundance of most
35 genera of sulfate-reducing bacteria (SRB), while stimulating the abundance of Se-tolerant
36 bacteria, which played an important role in Se(IV) reduction. Moreover, it was observed
37 that lactate was the preferred electron donor, linking with the production, and subsequent
38 consumption of acetate. X-ray absorption spectroscopy (XAS) and high-resolution
39 transmission electron microscopy (HRTEM) revealed the reduction of Se(IV) forming
40 amorphous Se(0) nanospheres. In addition, HRTEM showed that the biogenic Se(0)
41 undergo a biotransformation to more stable crystalline forms, contributing to the
42 immobilization of Se in the case of HLW release. Additionally, the sulfide generated by
43 the activity of SRB reacted with Cu producing corrosion products (Cu_xS) on the surface
44 of the copper material.

45 **1_ Introduction**

46 One of the challenges associated with the nuclear power industry is the final disposal of
47 spent nuclear fuel, which is classified as high-level radioactive waste (HLW). Among the
48 radionuclides generated by the fission of uranium fuel, selenium-79 (^{79}Se) stand out being
49 considered a key mobile fission product of particular concern in the disposal of spent fuel
50 and HLW (Hassan et al., 2021, Abdalla et al., 2020). Moreover, the chemical toxicity of
51 Se depends on its oxidation state (Ruiz-Fresneda et al., 2020; Avendaño et al., 2016). At
52 higher oxidation states (IV and VI), selenium is more soluble and thus mobile, and it is
53 therefore considered toxic. At lower oxidation states (0 and $-II$) Se species are less soluble
54 and mobile, exhibiting scarce environmental impact. The emission of beta particles during
55 its radioactive decay and its half-life (3.77×10^5 years) makes ^{79}Se a key determinant of
56 the long-term radiological impact of a deep geological repository (DGR) (Jörg et al.,
57 2010; Bienvenu et al., 2007). Ionizing radiation emitted by the radioisotope ^{79}Se could
58 contribute to the abiotic reduction of this metalloid in an aqueous environment. Ionizing
59 radiation emitted by the radioisotope ^{79}Se could contribute to the abiotic reduction of this
60 metalloid in an aqueous environment. Le Caër (2011) reported that when a radiation beam
61 crosses a system containing sodium selenite dissolved in water, it can promote the
62 radiolysis of the water molecules, producing large amounts of reducing agents (e.g.,
63 hydrated electrons, hydrogen radicals, or hydroxyl radicals). These generated products
64 exhibit a reduction potential that could reduce Se(IV) to Se(0), thereby facilitating the
65 prevalence of higher selenium oxidation states (El-Batal et al., 2020). However, this study
66 aims to evaluate the role of microbial communities in bentonite to assess the potential
67 selenium bioreduction as an additional barrier in the event of a waste leakage. Since
68 ionizing radiation can affect various factors, including metal oxidation states, microbial

69 viability and diversity, and the corrosion of metal canisters, this research builds on provide
70 a baseline understanding of the processes occurring in the absence of radiation exposure.

71 The most widely accepted solution for the long-term confinement of the HLW is the DGR
72 whose main aim of the DGR is to maintain its integrity for hundreds to millions of years,
73 preventing the release of radionuclides into the environment, until their radiotoxicity
74 decays to natural levels (WNA, 2024, Batandjieva et al., 2009). DGR's multi-barrier
75 system design consists of corrosion-resistant metal canisters surrounded by a backfill and
76 sealing material, and placed several hundred meters into a stable geological formation.

77 The material of DGR barriers depends on each country's approach. For example, Canada,
78 Korea, Sweden, and Finland have proposed or implemented a copper coat for their metal
79 canisters (Hall et al., 2021). Furthermore, the proposed material for buffering and sealing
80 DGRs is compacted bentonite clay. This material offers mechanical support for canisters
81 and helps seal cracks. Compacted bentonite would prevent water infiltration from
82 reaching the canisters, and hinder radionuclides diffusion (Garcia-Romero et al., 2019).

83 In Spain, the suggested clay material for the DGR is the bentonite from "El Cortijo de
84 Archidona" (Almeria), also known as FEBEX (Huertas et al., 2021; Villar et al., 2006).

85 Due to the fact that bentonites are not sterile, the characterization of their microbial
86 community structure and composition is crucial to assess the repository's long-term
87 stability and safety (Meleshyn et al., 2011). While the conditions within the DGR may
88 not be optimal for microbial growth, the permeation of nutrient-containing pore water
89 from the host rock might promote the development and activity of microorganisms. The
90 latter may cause microbially induced corrosion (MIC) of the metal canisters, contribute
91 to the production of a gaseous phase, alter inherent properties of the bentonite, and
92 influence mobilization or immobilization of radionuclides (Ruiz-Fresneda et al., 2023).

93 After the repository sealing, an anoxic environment could prevail due to microbial activity
94 and canister corrosion (Payer et al., 2019; Keech et al., 2014). Once this environment is
95 established, anaerobic bacteria such as sulfate-reducing bacteria (SRB) and iron-reducing
96 bacteria (IRB) might remain active negatively affecting the DGR barriers through
97 processes such as MIC promotion, bentonite illitization, and sulfate reduction to sulfide
98 (main canister corrosion agent) (Morales-Hidalgo et al., 2024; Pentráková et al., 2013;
99 Liu et al., 2012; Shelobolina et al., 2003). The sulfide generated by SRB, will be favored
100 in areas further from the canisters, where lower temperatures and greater water activity
101 could prevail (Bengtsson and Pedersen, 2017). This biogenic sulfide may diffuse through
102 the bentonite material and reach the canister surface causing corrosion and the release of
103 HLW in the worst-case scenario.

104 Despite the compaction of bentonite into blocks within the DGR, which should inhibit
105 microbial activity, all possible failure scenarios must be considered for safety
106 assessments. Therefore, the present work simulated a “worst-case” scenario where a
107 release of selenium could take place, and the bentonite turns to slurry form enriched with
108 nutrients. To assess the impact of microorganisms’ activity in the Spanish bentonite,
109 optimal growth conditions were established by setting up bentonite slurry microcosms
110 amended with electron donors (acetate and lactate) and acceptor (sulfate). Furthermore,
111 the addition of 2 mM of Se(IV) acted as a natural analogue of ⁷⁹Se present in the HLW.

112 Hence, this study aimed to examine, using an interdisciplinary approach, the impact of
113 Se(IV) amendment on the microbial communities of Spanish bentonite and the
114 subsequent effect of these microorganisms on the geochemical evolution of the
115 microcosms during one year of anoxic incubation. Additionally, an early-stage
116 characterization of microbial-mediated selenium reduction products and copper corrosion
117 was conducted. Our findings highlight the potential impact of bentonite microorganisms,

118 thriving under nutrient influx from water filtration on the clay and long-term release of
119 HLW from DGR.

120 **2_ Materials and methods**

121 **2.1_ Sampling of bentonite clay**

122 Bentonite was aseptically collected from “El Cortijo de Archidona” site (Spain),
123 disaggregated, dried, and ground as specified in Martinez-Moreno et al. (2023). The
124 chemical composition of the natural bentonite was previously published in Martinez-
125 Moreno et al. (2024b) and detailed in **Supplementary Table S1**.

126 **2.2_ Anoxic microcosms’ assembly: specifications, content, and sampling**

127 Microcosms were set up following the procedure stated by Martinez-Moreno et al.
128 (2024b) with some modifications. Briefly, 50 g of ground bentonite were deposited into
129 sterilized borosilicate glass bottles. Since bentonite is difficult to completely sterilize by
130 means of heat treatments (Martinez-Moreno et al., 2024a), one set of bentonite-containing
131 bottles (StB samples) were tyndallized (autoclave 110 °C for 45 min, 3 consecutive days)
132 in order to reduce the presence and activity of indigenous bentonite bacteria (hereafter
133 known as heat-shocked bentonites). To stimulate bacterial activity, some microcosms
134 (electron donors/acceptor-containing microcosms -eD-) were added with a final
135 concentration of 30 mM acetate, 10 mM lactate, and 20 mM sulfate. Bacterial consortium-
136 containing microcosms (BC) were spiked with a BPAS consortium (Martinez-Moreno et
137 al., 2024b), composed of four bacterial strains previously detected in the Spanish
138 bentonite (*Bacillus* sp. BII-C3, *Pseudomonas putida*, *Amycolatopsis ruanii*, and
139 *Stenotrophomonas bentonitica*), with an initial optical density (OD₆₀₀) of 0.4 each.
140 Additionally, three sterilized high-purity oxygen free copper mini canisters (Cu-mCan)

141 (Martinez-Moreno et al., 2024b) were deposited into each microcosm for corrosion
 142 studies.

143 All the microcosms were supplemented with a final concentration of 2 mM sodium
 144 selenite [Se(IV)] for the interaction studies. For this propose, 1M of sodium selenite
 145 [Se(IV)] stock solution ($\text{Na}_2\text{SeO}_3 \cdot \text{Sigma-Aldrich}$) was prepared and sterilized by 0.22
 146 μm pore-sized filters.

147 Lastly, the final volume of the microcosms was calculated taking into account all the
 148 solutions added and were filled up to 230 mL with equilibrium water (EW) (elaboration
 149 procedure and chemical composition previously detailed in Martinez-Moreno et al,
 150 2024b).

151 In summary, different treatments were considered (sample code between brackets): i)
 152 bentonite or heat-shocked bentonite (B – StB); ii) acetate, lactate, and sulfate addition
 153 (eD); and iii) spiked with the BPAS consortium (BC). Selenium-free treatments were
 154 previously published in Martinez-Moreno et al. (2024b). Overall, 24 microcosms were
 155 prepared (8 treatments in triplicate) and anoxically incubated at 28°C for one year. Sample
 156 ID and the content of each treatment are summarized in **Table 1**.

157 **Table 1.**
 158 Sample ID and content of the different microcosms. All microcosms contained each three Cu mini-canisters
 159 (Cu-mCan) and 2 mM of Se(IV). The microcosms were set up in triplicate and anoxically incubated at 28
 160 °C. ALS: acetate, lactate and sulfate concentrations; BPAS: bacterial consortium; +: presence; -: absence.

Sample ID	Bentonite	ALS (mM)	BPAS	Se(IV)
B.eD.Se.BC	Non-heat-shocked	30:10:20	+	2 mM
B.Se.BC		-	+	
B.eD.Se		30:10:20	-	
B.Se		-	-	
StB.eD.Se.BC	Heat-shocked	30:10:20	+	
StB.Se.BC		-	+	
StB.eD.Se		30:10:20	-	
StB.Se		-	-	

161 Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se:
 162 2 mM of Se(IV), BC: spiked with BPAS consortium.

163

164 The sampling procedure from the different microcosms for further analyses was carried
165 out within an anaerobic chamber. The step-by-step process of the experimental setup and
166 sampling is shown in **Figure 1**.

167 << INSERT FIGURE 1 >>

Figure 1. Step-by-step of the anaerobic microcosm's assembly, specifications, content, and sampling.

168

169 **2.3_ Geochemical analyses**

170 To determine the changes in physicochemical parameters, the supernatant (liquid phase)
171 was collected from each microcosm within an anaerobic chamber. Samples were
172 measured at 45, 90, 135, and 365 days of incubation, previously filtered through a 0.22
173 µm filter. pH measurements were performed by Advanced Digital Handheld Portable
174 Meter HQ40D from Hach, previously calibrated with commercially available reference
175 solutions (pH 4.00 and 7.00).

176 The percentage of acetate, lactate and sulfate from the amended treatments (eD) was
177 measured by high pressure ion chromatography (HPIC) by a 925 Eco ICE Methrom
178 Hispania (Herisau, Switzerland). Measurement specifications were followed in detail
179 according to Martinez-Moreno et al. (2024b).

180 **2.4_ Microbial DNA extraction, sequencing, and bioinformatic analyses**

181 Total DNA was extracted from the bentonite of the microcosms after 45 days of
182 incubation according to the protocol detailed by Povedano-Priego et al. (2021). The
183 concentration of the obtained DNA was measured using the Qubit 3.0 Fluorometer (Life
184 Technology, Invitrogen™).

185 For library preparation, the steps were followed as described in detail by Martinez-
186 Moreno et al. (2024b). The obtained FastQ files were processed and normalized using R
187 4.2.1 software (R Core Team, 2022). The assignment of phylotypes to taxonomic
188 affiliation was performed using a Bayesian classification (80% pseudo-bootstrap
189 threshold). Explicet 2.10.5 software (Robertson et al., 2013) was employed for the
190 calculation of the relative abundance and alpha diversity indices. A matrix based on the
191 Bray-Curtis algorithm through principal coordinate analysis (PCoA) was constructed to
192 evaluate the similarity between samples at the genus level using Past4 software (Hammer
193 and Harper, 2001). Additionally, a heatmap was generated to visualize the operational
194 taxonomic units (OTUs) causing dissimilarity between samples.

195 **2.5_ Characterization of selenium reduction products (SeRPs)**

196 **2.5.1_ Microscopic analyses**

197 Samples from the interface bentonite:EW (**Figure 1**) were recovered from the
198 microcosms at 45 days of incubation and analyzed by ultra-high resolution transmission
199 electron microscope and high-angle annular dark-field (HRTEM-HAADF) FEI TITAN
200 G2. One mL of each homogenized interface was centrifugated at 10,000 g for 5 min and
201 prepared as thin sections for microscopy analyses after contrasting with osmium for the
202 visualization of the bacterial cells. Afterwards, selenium reduction products (SeRPs) were
203 characterized by selected-area electron diffraction (SAED) and high-resolution
204 transmission electron microscopy (HRTEM) combined with fast Fourier transform (FFT).

205 **2.5.2_ X-ray absorption spectroscopy (XAS) analysis**

206 The interface from the microcosms were dried and powered for XAS analysis. The
207 powders were spread and pressed onto Kapton tape (10 mm × 3 mm area) and sealed

208 between additional layers. Selenium references, including sodium selenate [Se(VI)],
209 sodium selenite [Se(IV)], and elemental selenium [Se(0)] foil, were prepared following
210 the method described by Ruiz-Fresneda et al. (2020).

211 Se K-edge XAS data collected in fluorescence mode for the samples were obtained at the
212 SOLEIL synchrotron's MARS beamline (France). This bending magnet beamline is
213 dedicated to the multi-analyses of radioactive materials (Sitaud et al., 2012). Detailed
214 information about XAS measurements is specified in **section 1.1** of the **Supplementary**
215 **Material**.

216 **2.6_ Copper surface characterization**

217 The Cu-mCans were recovered from the microcosms after 45 days of anoxic incubation,
218 unscrewed, and left to dry under an anoxic atmosphere. Once dried, the bases of the Cu-
219 mCans were coated with carbon using an EMITECH K975X evaporator. Then the
220 samples were analyzed by high-resolution scanning electron microscopy (HRSEM) using
221 an AURIGA Carl Zeiss SMT microscope, coupled with qualitative and quantitative
222 energy dispersive X-ray (EDX) microanalysis.

223 In addition, chemical analysis using X-ray photoelectron spectroscopy (XPS) was
224 conducted on the surface of the Cu-mCan lids (with no previous preparation) using a
225 Kratos AXIS Supra photoelectron spectrometer following the procedure described by
226 Martinez-Moreno et al. (2024b). The spectrum peaks were fitted by CasaXPS 2.3.22
227 software (Fairley, 2019). All binding energies were offset by referencing the adventitious
228 carbon C1s peak (285 eV). No further constraints were applied to the initial binding
229 energy values.

230 3_ Results

231 3.1_ Visual transformations of the microcosms over one-year anoxic incubation

232 The color transformation occurring in the microcosms throughout the incubation time is
233 shown in **Figure 2**. After 45 days, microcosms containing BPAS consortium (BC) showed
234 a thin layer of reddish coloration at the bentonite:supernatant interface as well as
235 accumulations in some of the samples in the bentonite phase (red arrows, **Figure 2**). These
236 red products shifted, or were masked, by gray-black hues over the incubation period. The
237 formation of small fissures in the bentonite phase of the microcosms (white arrows,
238 **Figure 2**) and a characteristic smell of rotten eggs could be related to the gas generation.

239 << INSERT FIGURE 2 >>

Figure 2. Microcosms evolution over time of study (0, 45, 90, 135, and 365 days of incubation. Red arrows: precipitates attributed to Se(IV) reduction to Se(0). White arrows: fissures attributed to gas formation. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

240 The addition of the BPAS consortium (BC) in the microcosms stimulated the presence of
241 reddish precipitates observed in the early stages (45 days) of incubation in the samples.
242 Moreover, the addition of electron donors/acceptor (eD) also enhanced this color
243 transformation, while heat-shocked bentonite treatment slowed it. Interestingly, in the
244 sample of heat-soaked bentonite with Se (StB.Se), no visual color change was observed
245 throughout the year of incubation (**Figure 2**).

246 3.2_ Dynamics of geochemistry over one-year of incubation

247 The pH values in the microcosms exhibited alkaline to neutral values at the initial time
248 (t. 0) gradually decreasing and stabilizing around 7.50 – 8.50 (depending on the treatment)
249 after one year of incubation (t. 365) (**Supplementary Figure S2**).

250 The percentage of lactate, acetate, and sulfate in the amended microcosms (eD) is shown
251 in **Figure 3**. Lactate showed a fast consumption (**Figure 3A**). The depletion of this
252 electron donor was particularly evident in the microcosms spiked with the BPAS
253 consortium (B.eD.Se.BC), where total lactate depletion occurred within the first 45 days
254 of incubation. This process exhibited a gradual slowdown in the absence of the BPAS
255 consortium (B.eD.Se), followed by the heat-shocked bentonite microcosms with BPAS
256 consortium (StB.eD.Se.BC), and lastly, those with heat-shocked bentonite (StB.eD.Se).
257 In these last samples, complete lactate consumption was achieved before one-year
258 incubation. As expected, lactate consumption exhibited a close correlation with acetate
259 production in the microcosms (**Figure 3B**). The rise in acetate content, along with
260 subsequent consumption, was closely linked to lactate evolution, with the faster processes
261 observed in the B.eD.Se.BC sample, followed by B.eD.Se, StB.eD.Se.BC, and
262 StB.eD.Se.BC. It is noteworthy that, even after a year of incubation, acetate was not
263 completely consumed in any of the treatments. Only the B.eD.Se.BC treatment
264 approached complete acetate consumption before the end of the incubation year while, in
265 the heat-shocked samples, the consumption of acetate was poorly detected
266 (StB.eD.Se.BC) or only produced (StB.eD.Se). In the case of sulfate (**Figure 3C**), its
267 consumption was most pronounced in the non-heat-shocked samples containing the
268 BPAS consortium. In the samples B.eD.Se, StB.eD.Se.BC, and StB.eD.Se, sulfate content
269 remained relatively constant throughout the incubation time, slightly dropping after one
270 year of incubation (consumption of \approx 80%, 50%, 10%, respectively).

271 << INSERT FIGURE 3 >>

Figure 3. Variations in the initial concentration of 10 mM lactate (A), 30 mM acetate (B), and 20 mM sulfate (C) of the electron donors/acceptor-amended treatments. Data include mean values along with standard deviations, calculated from each independent microcosm. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

272 In summary, the progression of lactate consumption was intricately associated with
273 acetate production and, to a lesser extent, to sulfate consumption. Furthermore, the
274 addition of the BPAS consortium in the microcosms stimulated the mentioned processes,
275 whereas tyndallization (heat-shock) of the bentonite hindered them.

276 **3.3_ Characterization of the microcosm's interface precipitates**

277 As showed in **Figure 2**, the reddish coloration was more pronounced and heterogeneously
278 distributed in the samples spiked with the BPAS consortium (BC) after 45 days of
279 incubation. In contrast, the process was slower in the absence of the BPAS consortium.
280 For this reason, this early-stage of incubation was selected for the characterization of the
281 resulting products.

282 **3.3.1_ Microscopic and spectroscopic characterization**

283 The study of the Se reduction products, morphology and elemental composition was
284 determined by HAADF-HRTEM coupled with EDX maps, and their crystalline structure
285 was achieved by SAED and lattice spacing measured using HRTEM. No red layer in the
286 interface bentonite:supernatant at 45-days incubation was observed in the microcosms
287 lacking the BPAS consortium (**Figure 2**). In these samples, only bentonite was detected,
288 as shown by the Si, Al, Fe, Mg signals in the microanalyses and the elemental distribution
289 in the EDX maps (**Supplementary Figure S3**).

290 In the consortium-containing samples (BC samples), thin sections of the interface layer
291 showed the presence of electron-dense precipitates of different sizes, shapes and locations
292 (**Figure 4**). EDX analyses confirmed that these electron-dense accumulates corresponded
293 to the formation of Se species (selenium reduction products, SeRPs). As indicated by the
294 phosphorous (P) signal in the EDX maps (**Figure 4C**), all BC-treatments showed

295 intracellular selenium nanostructures or were found near cellular biomass with a rounded
296 shape. SAED analysis (**Figure 4B**) confirmed that these structures were amorphous (*a*-
297 Se). In addition, extracellular SeRPs with different morphology and size were found in
298 all samples (**Figure 4**). The elemental mapping of these electron-dense precipitates
299 (**Figure 4C**) in each sample showed a composition of Se and sulfur (S). The SAED
300 patterns (**Figure 4B**) indicated the crystallinity of these precipitates, while the lattice
301 spacing observed at high resolution (**Figure 4D**) showed values of 0.30 nm and 0.37 nm,
302 corresponding to the (101) and (100) plane of trigonal selenium (*t*-Se), respectively; and
303 0.38 nm corresponding to (211) plane of monoclinic selenium (*m*-Se).

304 << INSERT FIGURE 4 >>

Figure 4. HRTEM micrographs of the SeRPs from the BC microcosms at 45 days of incubation. A) HAADF images, B) associated SAED patterns, C) EDX maps showing the elemental distribution of Si, P, Se, Fe, and S; D) high resolution images and d-spacing. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

305 3.3.2_ X-ray absorption (XAS) analysis

306 X-ray absorption spectroscopy (XAS) spectra of the SeRPs in the Se-containing
307 treatments (B.Se.BC, B.eD.Se.BC, StB.Se.BC, and StB.eD.Se.BC) provided structural
308 data on the oxidation state of Se and local coordination of this element within the studied
309 treatments. The X-ray absorption near-edge structure (XANES) region revealed that Se(0)
310 dominated the local coordination of Se (**Figure 5**). This indicated that the whole amount
311 of reduced Se in solid form consisted of Se in the zero-valent oxidation state. The
312 extended X-ray absorption fine structure (EXAFS) spectra of the SeRPs and Se foil (Se⁰
313 reference compound) including their corresponding Fourier transforms (FT) and fit
314 parameters of the calculated spectra are shown in **Figure 5** and **Table 2**, respectively. FT
315 peak distances were uncorrected for scattering phase shift ($R + \Delta R$). The EXAFS fit

316 spectra analysis of the four samples revealed a single Se-Se coordination shell, with N
 317 values ranging from 1.9 to 2.7 and a bond distance of approximately $2.33\text{--}2.35 \pm$
 318 $0.0013\text{--}0.0026 \text{ \AA}$ (**Table 1**). The calculated bond distances were similar to those
 319 attributed to amorphous Se according to the literature ($2.33\text{--}2.34 \text{ \AA}$) (Eswayah et al.,
 320 2017; Ruiz-Fresneda et al., 2020; Povedano-Priego et al., 2023). The bond distance of the
 321 second Se–Se coordination shell ($3.63\text{--}3.70 \pm 0.02 \text{ \AA}$) would give insights on the
 322 crystalline Se, as was indicated by Eswayah et al. (2017). However, after several attempts
 323 failed due to the small amplitude of this peak in the four studied samples.

324 **Table 2.** EXAFS structural parameters of the Se foil and the selenium reduction products (SeRPs) from
 325 the selected Se-amended microcosms after 45 days of incubation.

326 ^aErrors in coordination numbers are $\pm 25\%$ and standard deviations as estimated by EXAFSPAK. ^bErrors
 327 in distance are $\pm 0.02 \text{ \AA}$. ^cDebye–Waller factor. Glossary: B: bentonite, StB: heat-shocked bentonite, eD:
 328 addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

329

330 << INSERT FIGURE 5 >>

Figure 5. (A) XANES spectra of Se reference compounds Na_2SeO_3 (Se^{IV}), Se foil (Se^0), and Na_2SeO_4 (Se^{VI}), and SeRPs from B.eD.Se.BC, B.Se.BC, StB.eD.Se.BC and StB.Se.BC samples after 45 days of incubation. (B) EXAFS spectra of Se foil and SeRPs from the microcosms and (C) their corresponding FT. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

331 3.4_ Analysis of microbial communities

Sample	Shell	N^a	R^b [\AA]	σ^{2c} [\AA^2]	ΔE [eV]
Se foil	Se–Se	3.2 ± 0.2	2.37	0.0040	–10.10
B.eD.Se.BC	Se–Se	1.9 ± 0.2	2.33	0.0013	–13.70
B.Se.BC	Se–Se	2.7 ± 0.2	2.34	0.0026	–12.37
StB.eD.Se.BC	Se–Se	2.4 ± 0.1	2.35	0.0024	–11.10
StB.Se.BC	Se–Se	2.5 ± 0.1	2.35	0.0025	–11.60

332 Total DNA was extracted and sequenced from the bentonite of the microcosms after 45
 333 days of incubation (where the microcosms exhibited greater visual heterogeneity between

334 treatments). The samples B.Se.BC_R1 and StB.eD.Se_R1 were not included in this study
335 due to their failure during the DNA extraction step. A total of 353 phlotypes were
336 annotated in 12 phyla (98.34%) being the most abundant Pseudomonadota (68.00%) and
337 Bacillota (26.58%) (**Supplementary Table S2**). At the genus level, 135 OTUs were
338 annotated as *Pseudomonas* (40.30%), *Stenotrophomonas* (12.52%), *Symbiobacterium*
339 (12.43%), unclassified Clostridia (6.93%), unclassified Rhodospirillales (6.00%), and
340 *Pseudoalteromonas* (2.79%) showing the higher relative abundance in the pool of
341 samples (**Supplementary Table S3**).

342 The principal coordinate analysis (PCoA) represented the dissimilarity between the
343 samples spiked with (BC samples) and without the BPAS consortium. Within BC samples
344 (PCoA with this subset of samples, **Figure 6**), no difference was observed between heat-
345 shocked (StB) and non-heat-shocked (B) bentonite. However, in the treatments without
346 BPAS consortium, the dissimilarity between heat-shocked bentonite (StB) and non-heat-
347 shocked bentonite (B) was evidenced (**Figure 6**).

348 << INSERT FIGURE 6 >>

Figure 6. Principal coordinate analysis (PCoA) comparing the bacterial communities at genus level from Se(IV)-containing microcosms after 45 days of incubation. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

349 To identify the OTUs involved in the dissimilarity between samples, a heatmap was
350 constructed (**Figure 7**). In the absence of BPAS consortium in the heat-shocked bentonite
351 samples (StB cluster, green bracket **Figure 7**), bacterial diversity significantly diminished
352 compared to that of the other treatments. The marked differentiation in these treatments
353 was due to the abundance of *Pseudomonas*, *Symbiobacterium*, *Stenotrophomonas*, and
354 unclassified Clostridia. Concerning the samples spiked with the BPAS consortium (BC
355 cluster, purple bracket **Figure 7**), the observed group was in line with expectations, as it

356 was attributed to the bacterial genera inherent to the consortium, such as *Pseudomonas*,
357 *Stenotrophomonas* and, to a lesser extent, *Amycolatopsis*. In the cluster of non-heat-
358 shocked bentonite (B cluster, brown bracket **Figure 7**), the abundance of unclassified
359 Clostridia, unclassified Rhodospirillales, *Pseudoalteromonas*, *Noviherbaspirillum*, and,
360 to a lesser extent, *Pseudomonas*, marked the dissimilarity from the other treatments.

361 << INSERT FIGURE 7 >>

Figure 7. Heatmap showing the bacterial relative abundance and clustering at genus-level of the microcosms after 45 days of incubation. Cut-off: 1.33% based on the maximums. The warmer the color, the greater is the relative abundance. The colored lines in X axis represent the clustering observed in PCoA. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

362 The alpha-diversity indices of the treatments (duplicates in B.Se.BC and StB.eD.Se) at
363 the genus level are shown in **Supplementary Table S4**. The richness index (Sobs) of the
364 non-heat-shocked samples (B samples) showed higher values in the treatments without
365 BPAS consortium (BC), particularly higher values in the absence of electron
366 donors/acceptor (B.Se > B.eD.Se > B.eD.Se.BC > B.Se.BC). In contrast, the samples with
367 heat-shocked bentonite (StB samples) exhibited a different trend. The richness indices
368 were higher in the treatment with electron donors/acceptor, followed by samples spiked
369 with the BPAS consortium, the treatment with selenium only, and finally, the samples
370 amended with electron donors/acceptor and the BPAS consortium (StB.eD.Se >
371 StB.Se.BC > StB.Se > StB.eD.Se.BC). Regarding the diversity values, the treatments
372 with non-heat-shocked bentonite (B.Se) had the highest values, followed by the samples
373 amended with electron donors/acceptor (B.eD.Se), whose ShannonH and SimpsonD
374 values were greater than 3 and close to 1, respectively. The remaining treatments
375 exhibited lower values, with ShannonH lower than 3 (1.37–2.36) and SimpsonD further

376 from 1 (0.39–0.64) (**Supplementary Table S4**). These values suggested reduced
377 diversity in the latter samples and the potential genus dominance in the microcosms.

378 The relative abundance of OTUs in each treatment in triplicate (duplicates in B.Se.BC
379 and StB.eD.Se) are shown in **Figure 8** and **Supplementary Table S3**. Overall,
380 *Pseudomonas* emerged as the predominant OTU in this study across the sample pool
381 (40.30% of the total relative abundance). Its dominance was particularly pronounced in
382 samples spiked with the BPAS consortium (BC samples), where it constituted the taxa
383 dominance in the community, ranging from 77.20% of relative abundance in sample
384 B.Se.BC to 48.18% in sample B.eD.Se.BC. Within this subset of samples,
385 *Stenotrophomonas* (member of the BPAS consortium) emerged as the second most
386 represented genus, with relative abundance ranging from 34.47% in sample B.eD.Se.BC
387 to 10.79% in sample B.Se.BC. The relative abundance of *Amycolatopsis* in the BC
388 treatments was comparatively lower than the other members of the consortium, ranging
389 between 4.63% in sample B.eD.Se.BC and 0.38% in sample B.Se.BC. Conversely,
390 *Bacillus* was absent possibly for sequencing limitations.

391 << INSERT FIGURE 8 >>

Figure 8. Bacterial community structure within the bentonite microcosms showing the OTU relative abundance (averages of biological replicates) at genus level. Cut-off: 0.20%. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

392 Regarding samples lacking the BPAS consortium (B.eD.Se and B.Se), the OTU with the
393 highest relative abundance was unclassified Rhodospirillales (23.18% and 20.65%,
394 respectively). This group was followed by unclassified Clostridia (23.14%),
395 *Desulfosporosinus* (9.05%), and *Pseudomonas* (8.89%) in sample B.eD.Se. In the case of
396 B.Se, the following most abundant OTUs were *Pseudoalteromonas* (19.11%),
397 unclassified Clostridia (13.94%), and *Noviherbaspirillum* (8.55%). The bacterial

398 diversity present in the microcosms was not greatly affected by the addition of electron
399 donors/acceptor. However, there were some differences observed. Sample B.eD.Se had a
400 higher presence of *Desulfosporosinus* and *Alkalihalobacillus* (9.05% and 8.52%,
401 respectively) compared to B.Se (1.07% and 0.21%). On the other hand,
402 *Pseudoalteromonas* was more prominent in sample B.Se (19.11%) than in B.eD.Se
403 (0.21%).

404 **3.5_ Microscopic and spectroscopic characterization of the Cu-mCan's surface**

405 To evaluate the impact of microbial communities and the presence of selenium on the
406 copper material after 45-day incubation, one Cu-mCan was collected from each
407 microcosm. The surfaces underwent microscopic and spectroscopic analyses by HRSEM,
408 EDX, and XPS. The copper surface characterization was focused on Cu-mCan recovered
409 from non-heat-shocked bentonite microcosms.

410 **Figure 9** presents Cu-mCan and HRSEM images before (t. 0) and after a 45-day
411 anaerobic incubation. Visual differences in coloration were observed compared to the
412 initial state of Cu-mCan (t. 0). It is worthy to note that the bentonite adhering to the copper
413 surface was intentionally retained to prevent potential displacement of copper alteration
414 related compounds. The sample B.eD.Se.BC showed a significant shift towards dark
415 gray-black coloration (**Figure 9B**). That hue was also observed in some regions of
416 B.Se.BC and B.eD.Se (**Figure 9C and 9D**). At the microscopic level, the post-incubation
417 samples exhibited heterogeneous topography (second column, **Figure 9**), in contrast to
418 the uniformity observed in the t. 0 sample. Cu-mCan at t. 0 only showed the presence of
419 Cu and C in the EDX maps. The presence of C was attributable to the carbon metallization
420 process during sample preparation for HRSEM. In addition, the subsequent columns of
421 **Figure 9** present elemental EDX maps that illustrate the distribution of iron (Fe),

422 aluminum (Al), sulfur (S), silicon (Si), oxygen (O), copper (Cu), and carbon (C) in
423 selected areas. The Si (pink), Fe (yellow), and Al (cyan) signals were assigned to the
424 residual bentonite adhered to the Cu surface. The signal of O (red) was mainly related to
425 Si (Si-O bonds), indicating its presence within the bentonite. Additionally, to a lesser
426 extent, O was detected on the surface of the copper.

427 Sample B.eD.Se.BC showed sulfur accumulates, which may be associated with copper
428 corrosion (Cu_xS) (**Figure 9B**) linked with the consumption of sulfate ($\approx 8\%$) (**Figure 3**).
429 The sample B.Se.BC also showed S-accumulates (**Figure 9C**). However, in samples
430 B.eD.Se and B.Se no sulfur was observed, and the detection of O was mainly related to
431 the presence of bentonite (**Figure 9E** and **9D**). On the surface of these Cu-mCan, some
432 "patches" were observed. Upon closer examination at higher magnification, these patches
433 appeared to indicate some delamination on the copper surface, potentially exposing bare
434 Cu to new oxidation processes (**Figure 9D**, orange square). No Se signal was detected on
435 the copper surface by EDX.

436 << INSERT FIGURE 9 >>

Figure 9. Visual images (first column), electron images, and EDX maps of the Cu-mCan surface before experimental set-up (A) and after 45-days incubation from treatments B.eD.Se.BC (B), B.Se.BC (C), B.eD.Se (D), and B.Se (E). EDX maps show the distribution of Cu (orange), Fe (yellow), S (dark blue), Si (pink), Al (cyan), O (red), and C (green) in the studied area. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

437 Additionally, XPS wide scans of the copper lids indicated the presence of Cu, C, O, Si,
438 Na, Fe, Mg, and Al on all Cu-mCan surfaces. These elements (excluding copper) were
439 attributed to the residual bentonite present on the copper surfaces. The high-resolution
440 scans by XPS (**Figure 10**) of the Cu 2p region revealed the presence of elemental copper
441 on all sample surfaces, indicated by the peaks at 932.6 eV and 952.3 eV (Wagner, 1979;
442 Mansour, 1994). Additional peaks at 933.6 eV and 953.70 eV evidenced the presence of

443 CuO on all the samples (Ghijsen et al., 1988; Jolley et al., 1989), and the satellite peaks
444 at around 942 eV and 962 eV were assigned to the presence of Cu²⁺ (Biesinger, 2017). It
445 should be noted that, although the intensity of the Se signal in some of the scanned copper
446 lids was low, given the low concentration expected for this element on the surface of the
447 copper, the presence of Se was still evident in all the samples. The high-resolution scans
448 of the Se 3d region (**Figure 10**) showed two peaks in all samples: a signal around 58.2
449 eV, which has been previously reported for Se(IV) from Na₂SeO₃ (Moulder et al., 1992),
450 and another around 54.8 eV, usually attributed to Se(0) (Weser et al., 1977). The presence
451 of both peaks indicates the reduction of Se(IV) to Se(0). In the microcosms with electron
452 donors, the peaks of both Se(IV) and Se(0) could be clearly appreciated, but the intensity
453 of Se(IV) and Se(0) are different. While the Se(IV) was evident, the peak corresponding
454 to Se(0) was much more intense and defined, with the possibility of distinguishing the
455 contributions from Se 3d_{3/2} and Se 3d_{5/2} separately. For example, in sample B.eD.Se the
456 signal for Se(0) 3d_{3/2} appeared at 55.10 eV, and the peak for Se(0) 3d_{5/2} appeared at 54.24
457 eV. The separation between these two signals at 0.86 eV is in agreement with previously
458 reported XPS Se spectra (Moulder et al., 1992).

459 << INSERT FIGURE 10 >>

Figure 10. High-resolution XPS spectra of the regions Cu 2p (977 eV – 922 eV) and Se 3d (70 eV – 45 eV) of the copper surface recovered from the microcosms after 45-days of incubation. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

460 **4_ Discussion**

461 **4.1_ Microcosms evolution and dynamics of the water geochemistry**

462 The visual color changes of the microcosms over one-year of anoxic incubation showed
463 a thin layer of reddish coloration at the interface and accumulations of reddish spots in
464 the bentonite phase. This coloration is an indicator of the reduction of selenite [Se(IV)]

465 to elemental selenium [Se(0)] (Povedano-Priego et al., 2023; Ruiz-Fresneda et al., 2018).
466 These red Se(0) reduction products (SeRPs) shifted, or were masked by gray-black hues
467 over the incubation period in the observed microcosms. This shift towards darker hues
468 could be attributed to factors such as a transition in the Se allotropy of the SeRPs to more
469 stable forms, or to the generation of the bacterial-produced H₂S, which can react with iron
470 resulting in black precipitates of reduced iron species (Ruiz-Fresneda et al., 2023b;
471 Miettinen et al., 2022; Matschiavelli et al., 2019). The formation of small fissures in the
472 bentonite phase of the microcosms (white arrows, **Figure 2**) and the characteristic smell
473 of rotten eggs could be related to the gas generation (e.g., sulfide) mediated by bacterial
474 activity (He et al., 2011). However, further investigations will need to determine the gases
475 composition. It should be noted that the addition of the BPAS consortium (BC) in the
476 microcosms promoted Se reduction process, with reddish coloration observed at 45-days
477 incubation. The addition of electron donors/acceptor (eD) also enhanced the reduction
478 process, while bentonite heat-shock slowed down this color shift. Regarding the sample
479 of heat-shocked bentonite with Se (StB.Se), no visual color transition was apparent
480 throughout the year of incubation indicating no Se(IV) reduction either alteration to
481 darker hues in the bentonite. Since tyndallization is not the most effective process for
482 bentonite sterilization (Martinez-Moreno et al., 2024a), it was observed that, under the
483 studied conditions, this process can influence bacterial activity.

484 Regarding the dynamic in the water geochemistry, the pH values in the microcosms over
485 the course of incubation exhibited a tendency to become less alkaline, stabilizing within
486 neutral values. This trend in the pH values was in accordance with previous studies
487 without Se conducted by Martinez-Moreno et al. (2024b) where indicated that the
488 observed trend in the pH might be related to the ionic exchange occurring between
489 bentonite and the different solvents (e.g., acetate, lactate, sulfate, and the EW) and/or the

490 gas generation induced by the microbial communities' metabolisms. However, the
491 addition of 2 mM of Se(IV) did not impact the pH trend of the microcosms. The
492 progression of lactate consumption was intricately associated with acetate production
493 and, to a lesser extent, to sulfate consumption. Furthermore, the addition of the BPAS
494 consortium in the microcosms stimulated the consumption and/or production of lactate,
495 acetate and sulfate, whereas the heat-shock of the bentonite hindered them. This dynamic
496 was also observed in Martinez-Moreno et al. (2024b) with an important difference: the
497 presence of Se(IV) slowed the dynamics of the electron donors/acceptor. In the cited
498 work, the authors proposed that microorganisms could use lactate, without sulfate
499 reduction, through lactate fermentation pathway, which result in the production of acetate
500 and propionate, or by the incomplete oxidation of lactate to acetate coupled with the
501 reduction of iron (Park et al., 2024). Additionally, the decrease in sulfate content observed
502 in the treatment B.eD.Se.BC may be due to the sulfate reduction to hydrogen sulfide with
503 the concurrent incomplete oxidation of lactate to acetate through the reductive pathway
504 of acetyl-CoA (Matschiavelli et al., 2019). This last component formation could be related
505 to the fissures observed in the microcosms (**Figure 2**) and the rotten egg odor after
506 sampling (He et al., 2011). It is noteworthy that, even after a year of incubation, acetate
507 was not completely consumed in any of the treatments. Only the B.eD.Se.BC treatment
508 approached complete acetate depletion before the end of the incubation year. This finding
509 demonstrated the negative impact of the Se(IV) addition on the activity of the bacterial
510 community slowing the consumption/production of lactate, acetate, and sulfate compared
511 to Se(IV)-untreated microcosms described in Martinez-Moreno et al., 2024b.

512 The presence of Se(IV) at 2 mM can be toxic to some bacteria (Povedano-Priego et al.,
513 2023) and might reduce the bacterial activity within the microcosms. Moreover, this
514 metalloid could act as a competitor of sulfur as a final electron acceptor under anaerobic

515 condition (Staicu and Barton, 2021) associated to the slow consumption in the
516 microcosms (**Figure 3C**). This fact was observed comparing samples B.eD.Se.BC and
517 B.eD.Se, where sulfate consumption was lower in the presence of the BPAS consortium
518 showing the preference of those bacteria to use selenite as electron acceptor within a
519 detoxification process instead of sulfate. The presence of Se(IV) could lead to a
520 deceleration of biogeochemical processes, attributed to a reduction in bacterial diversity
521 and/or an extended adaptation period for selenium-tolerant microorganisms.

522 **4.2_ Effect of Se(IV) addition on the microbial communities in the early-incubation** 523 **stage.**

524 Comparing the relative abundances of the bacterial strains from the BPAS consortium
525 (*Bacillus*, *Pseudomonas*, *Amycolatopsis* and *Stenotrophomonas*) in this study with those
526 microcosms without Se(IV) from the previous study by Martinez-Moreno et al. (2024b),
527 several differences were noted. In the presence of Se(IV), *Pseudomonas* decreased from
528 69.53% to 48.18%, and from 78.88% to 77.20% in B.eD.Se.BC and B.Se.BC treatments,
529 respectively, compared to its counterpart without Se(IV). Conversely, in the heat-shocked
530 bentonite (StB) treatments, *Pseudomonas* increased from 57.73% to 62.80%, and from
531 54.59% to 73.44% in StB.eD.Se.BC and StB.Se.BC treatments, respectively, compared
532 with the samples without Se, which are detailed in Martinez-Moreno et al. (2024b). For
533 *Stenotrophomonas*, the addition of Se(IV) increased its abundance from 8.12% to 34.47%
534 in B.eD.Se.BC treatment compared to B.eD.BC, while it decreased from 34.96% to
535 13.24% in StB.Se.BC treatment compared to StB.BC. *Amycolatopsis* showed a higher
536 abundance (from 2.31% to 4.63%) in B.eD.Se.BC, and was detected in StB.Se.BC at
537 1.21%, compared to its absence in the Se(IV)-free counterpart.

538 The effect of Se(IV) addition in the microcosms appears to have a potentially negative or
539 positive impact on the BPAS consortium members depending on the treatment and the

540 bacterium. This effect is compensated by a balanced increase-decrease in *Pseudomonas*
541 and *Stenotrophomonas*, which were the dominant genera in the spiked microcosms.
542 *Pseudomonas* and *Stenotrophomonas* have demonstrated tolerance and the ability to
543 interact with Se(IV) by reducing it to Se(0) under anoxic conditions, resulting in the
544 formation of orange-reddish precipitates in form of selenium nanoparticles (SeNPs)
545 (Staicu et al., 2015; Ruiz-Fresneda et al., 2019). This process is linked to the presence of
546 red precipitates observed in the microcosms spiked with BPAS consortium at 45 days-
547 incubation (BC microcosms, **Figure 2**). The presence of the BPAS consortium stimulated
548 the Se(IV) reduction processes since in the unspiked treatments, the reddish precipitates
549 were less evidenced after 45 days of incubation (**Figure 2**). These bacteria have
550 demonstrated the capacity to metabolize acetate and/or lactate as a carbon source under
551 anoxic conditions (Freikowski et al., 2010; Sanchez-Castro et al., 2017) and could play
552 an important role in the lactate consumption and acetate production observed after 45
553 days of incubation in the BC treatments (**Figure 3**).

554 In samples lacking the BPAS consortium (B.eD.Se and B.Se), the presence of Se(IV)
555 caused shifts in bacterial community composition compared to its counterparts without
556 Se(IV) observed by Martinez-Moreno et al. (2024b). Following the results observed in
557 the mentioned study, the addition of Se(IV) eliminated the presence of *Desulfocurvibacter*
558 and *Desulfuromonas* in B.eD.Se and B.Se. Additionally, electron donors/acceptor had
559 minimal effect on bacterial diversity. Treatment B.eD.Se showed higher levels of
560 *Desulfosporosinus* and *Alkalihalobacillus* compared to B.Se. Conversely,
561 *Pseudoalteromonas* was more abundant in B.Se than in B.eD.Se.

562 *Desulfosporosinus* is a SRB capable of reducing sulfate via incomplete oxidation of
563 lactate resulting in the production of acetate and CO₂ (Stackebrandt et al., 1997), and may
564 be one of the bacteria involved in the slight production of acetate observed at 45-days

565 incubation in the B.eD.Se sample (**Figure 3**). In addition, their tolerance to the presence
566 of Se was previously demonstrated (Povedano-Priego et al., 2023; Aoyagi et al., 2021).
567 Additionally, *Alkhalibacillus* is a recently described genus with some members
568 facultative anaerobic and tolerant to heavy metals and metalloids such as chromium or
569 arsenic, which may be associated to the tolerance to Se(IV) (Patel and Gupta, 2020;
570 Gautam et al., 2022; Haghi et al., 2023). Regarding the heat-shocked bentonite treatments
571 without the BPAS consortium (StB.eD.Se and StB.Se), it is worth noting the high relative
572 abundance of *Symbiobacterium* (58.97% and 49.21%, respectively). This anaerobic
573 bacterium can grow in temperatures ranging from 40 to 70 °C (Shiratori-Takano et al.,
574 2014). Previous studies have identified this genus with an important relative abundance
575 in selenite-containing soils (Povedano-Priego et al., 2023; Aoyagi et al., 2021).

576 Overall, the presence of Se(IV) influences the bacterial communities within the
577 microcosms. Therefore, it is important to the explore these communities' role in the
578 interaction with this metalloid and its impact on Se speciation.

579 **4.3_ Characterization of the selenium reduction products (SeRPs)**

580 In order to elucidate the first steps of Se(IV) reduction, this section is focused, as
581 previously mentioned, on an early-stage (45-day incubation) characterization of the
582 selenium reduction products (SeRPs). At this incubation time, Se reduction was more
583 pronounced in the samples containing the BPAS consortium (BC samples). The ability of
584 the BPAS strains to reduce Se(IV) to Se(0) has been previously reported by Povedano-
585 Priego et al. (2023), where these bacteria promoted Se(IV) reduction resulting in the
586 production of nanoparticles of Se(0). Furthermore, the predominant bacteria in the BC-
587 containing microcosms of this study, *Pseudomonas* and *Stenotrophomonas*, were the most
588 abundant genera (**Figure 8**), which showed the ability to tolerate selenium through its

589 reduction to Se(0) nanoparticles (Staicu et al., 2015; Ruiz-Fresneda et al., 2019). In
590 addition to biotic factors, Se(IV) immobilization can be mediated by abiotic processes
591 (Wang et al., 2022). These abiotic processes include the adsorption of Se to the surface of
592 clay minerals, its reduction to elemental Se mediated by pyrite (FeS₂) oxidation, or via
593 structural Fe(II) present in smectites (Hoving et al., 2019; Kang et al., 2011; Breynaert et
594 al., 2010). In contrast, no abiotic reduction of selenite was detected in our study (sample
595 StB.Se, **Figure 2** and **Supplementary Fig. S1**).

596 Selenium can manifest in three different allotropes: amorphous (*a*-Se), monoclinic (*m*-
597 Se), and trigonal (*t*-Se), with *a*-Se considered thermodynamically unstable, while *t*-Se
598 represents the most stable allotrope. The structural analysis of the SeRPs in this study
599 unveiled that intracellular selenium precipitates, in form of nanoparticles (SeNPs),
600 exhibited amorphous allotropy. In contrast, extracellular electron-dense aggregates were
601 identified as SeRPs with monoclinic and trigonal allotropy, occasionally combined
602 (sample StB.Se.BC, **Figure 4**). Notably, the extracellular crystalline SeRPs demonstrated
603 a close association with sulfur (S), as evidenced in the EDX maps (**Figure 4C**). This
604 suggests that the reduction of Se(IV) to Se(0) might occur within the bacterial cells,
605 resulting in the formation of amorphous SeNPs (*a*-Se). Subsequently, these particles
606 could be released into the extracellular environment through mechanisms such as cellular
607 detoxification or lysis. Once in the extracellular space, the transformation process to *m*-
608 Se and, subsequently, to *t*-Se would take place. This finding aligned with those of previous
609 research involving complex saturated-bentonite microcosms under anaerobic conditions
610 (Povedano-Priego et al., 2023) and supported the mechanism proposed by Ruiz-Fresneda
611 et al. (2018). XAS was used to get insights into the local coordination of Se within the
612 samples (e.g., Se-Se bond distances) which help to determine the proportion of different
613 Se allotropes. The EXAFS fit analysis indicated that Se(0) is mainly present as amorphous

614 form where a Se-Se bond distance is observed, slightly lower than that of the trigonal Se
615 allotrope (Se-Se bond distance ≈ 2.37 Å). Since EXAFS is a bulk technique, no solid
616 structure evidence for the transformation of *a*-Se to *t*-Se was detected as suggested by
617 STEM/HAADF. However, low proportion of *t*-Se is not discarded in the studied samples.
618 Moreover, the presence of sulfate in the microcosms could mediate the interaction of the
619 SeNPs with S to form the observed selenium-sulfur precipitates in the extracellular
620 environment (Vogel et al., 2018). The Se-S interaction manifested as precipitates may
621 impede sulfate availability that could impact the corrosion of copper material. It is crucial
622 to explore not only its influence on bacterial communities but also how the presence of
623 Se(IV) in the microcosms might affect the microbial-influenced corrosion of Cu-mCan.

624 **4.4_ Chemistry of the copper surface**

625 The signal of Se(0) detected by XPS on the copper surface, showed its highest intensity
626 in the sample B.eD.Se.BC, indicating that the addition of electron donors/acceptor and
627 the BPAS consortium presented the highest effect in the Se reduction processes.
628 According to the literature, Cu_xS is expected to exhibit a Cu $2p_{3/2}$ peak around 932 eV
629 (**Figure 10**), although confirming the presence of this compound is challenging because
630 this signal would be overshadowed by the elemental Cu peaks (Gebhardt et al., 1986;
631 Romand et al., 1978). Nonetheless, as stated in Martinez-Moreno et al. (2024b), the
632 presence of S-aggregates observed by microscopy (**Figure 9**) on the copper surface of
633 treatment B.eD.Se.BC and B.Se.BC might be related to the presence of Cu_2S resulted
634 from microbial activity. Still, additional techniques should be carried out to better
635 understand the speciation of the S-aggregates.

636 The use of lactate as electron donor by SRB coupled to the reduction of sulfate to sulfide
637 under anoxic can occur via dissimilatory sulfate reduction pathway (Dou et al., 2020).

638 Moreover, Salehi Alaei et al. (2023) reported that the transformation of copper oxide
639 (Cu_2O) into copper sulfide (Cu_2S) through a chemical exchange with sulfide could be
640 thermodynamically possible. In addition, the formed HS^- , resulting from the SRB
641 metabolite H_2S , could migrate to the copper surface and react with it leading to the
642 formation of Cu_2S (Dou et al., 2020).

643 SRB could bond to copper surfaces creating a primary corrosion film of Cu_2S (Chen et
644 al., 2014). The creation of this Cu_2S film and the production extracellular polymeric
645 substances (EPS) by SRB can reduce the toxicity of copper (from the canister surface) to
646 this group of bacteria. Moreover, the presence of SRB has been proven to cause high-
647 purity copper corrosion in groundwater boreholes or MX-80 clay, leading to Cu_2S
648 precipitate formation (Johansson et al., 2017; Masurat et al., 2010). In a recent study,
649 Martinez-Moreno et al. (2023) demonstrated that the addition of acetate, lactate, and
650 sulfate in bentonite blocks compacted to a high dry density stimulated bacterial activity,
651 resulting in the formation of small corrosion compounds on the surface of copper disks.
652 In the present work, the high presence of water accelerated the corrosion processes and
653 the formation of precipitates was observed as early as 45 days of incubation.

654 **4_ Conclusions**

655 Visually, the addition of the BPAS consortium in the microcosms promoted the processes
656 of Se(IV) reduction to Se(0) , followed by the addition of electron donors/acceptor. In
657 contrast, the tyndallization (heat-shock) of the bentonite delayed these processes, with no
658 selenium reduction (red color) detected in the StB.Se sample over the year of incubation.
659 The addition of Se(IV) affected microbial-mediated biogeochemical process due to its
660 toxicity for some bacteria within the bentonite community. On the other hand, the
661 presence of nutrients such as lactate, acetate and sulfate stimulated microbial activity.

662 The presence of Se(IV) in the microcosms reduced the presence of some SRB, while the
663 electron donors/acceptor stimulated the growth of the Se-tolerant SRB
664 *Desulfosporosinus*. Additionally, Se-tolerance bacteria such as *Pseudomonas*,
665 *Stenotrophomonas*, and *Alkalibacillus* may contribute to the visual evolution in the
666 microcosms, indicated by the evidence of Se(IV) reduction to Se(0) over the year-long
667 anaerobic incubation period. In samples with electron donors/acceptor, lactate was the
668 main carbon source to be consumed, associated with acetate production and subsequent
669 consumption, and to a lesser extent, sulfate reduction. The dynamic rate of these processes
670 in the samples increased upon the nature of the biological of physicochemical parameter
671 in the following order: spiked BPAS consortium > amended electron donors/acceptor >
672 bentonite heat-shock.

673 The presence of the BPAS consortium revealed the reduction of selenium. The suggested
674 process started with the intracellular uptake and reduction of Se(IV) to Se(0) in form of
675 nanospheres with amorphous allotropy (*a*-Se), which in the extracellular space, led to the
676 transformation of extracellular crystalline structures, progressing from monoclinic
677 selenium (*m*-Se) and concluding with the most stable form of trigonal selenium (*t*-Se).

678 SRB oxidized lactate by using sulfate as final electron acceptor via dissimilatory pathway.
679 The resulting biogenic sulfide may participate in a chemical substitution with copper
680 oxides, which could be the result of the presence of oxygen that has been trapped in the
681 bentonite or the reduction of H₂O molecules, resulting in the formation of copper sulfide
682 (Cu₂S). Moreover, HS⁻ could react to copper surface and deal with the formation of Cu₂S
683 precipitates. However, further analysis are required to investigate the impact of Se(IV) on
684 copper biocorrosion.

685 Overall, the presence of Se(IV) impacted the bacterial communities inherent of the
686 Spanish bentonite by affecting their activity. Furthermore, the enrichment of selenium-
687 tolerant bacteria with the capacity to reduce Se(IV) to Se(0) could benefit the DGR safety
688 due to the immobilization of this metalloid, thus enhancing the repository's environmental
689 safety.

690 **Funding**

691 The present work was supported by the grant RTI2018–101548-B-I00 “ERDF A way of
692 making Europe” to MLM from the “Ministerio de Ciencia, Innovación y Universidades”
693 (Spanish Government). The project leading to this application has received funding from
694 the European Union’s Horizon 2020 research and innovation programme under grant
695 agreement No 847593 to MLM. ADM acknowledges funding from the UK Engineering
696 and Physical Sciences Research Council (EPSRC) DTP scholarship (project reference:
697 2748843).

698 **Acknowledgements**

699 The authors would like to thank Dr. F. Javier Huertas (IACT, Spain) for his invaluable
700 guidance and assistance in the collection of bentonites from El Cortijo de Archidona.
701 Rahul N. Doulatram Gamgaram and Maria del Carmen Contreras Morales (Institute of
702 Water Research, University of Granada) for conducting the HPIC measurements. Daniel
703 García-Muñoz Bautista-Cerro and Alicia González Segura (Centro de Instrumentación
704 Científica, University of Granada, Spain) for their support with sample preparation and
705 microscopy assistance, respectively.

706 **References**

- 707 Abdalla, Z. A. Y., Ismail, M. Y. A., Njoroge, E. G., Hlatshwayo, T. T., Wendler, E., & Malherbe, J. B. 2020.
708 Migration behaviour of selenium implanted into polycrystalline 3C–SiC. *Vacuum*, 175, 109235.
709 <https://doi.org/10.1016/j.vacuum.2020.109235>
- 710 Aoyagi, T., Mori, Y., Nanao, M., Matsuyama, Y., Sato, Y., Inaba, T., ... & Hori, T. 2021. Effective Se
711 reduction by lactate-stimulated indigenous microbial communities in excavated waste rocks. *J.*
712 *Hazard. Mater.*, 403, 123908. <https://doi.org/10.1016/j.jhazmat.2020.123908>
- 713 Avendaño, R., Chaves, N., Fuentes, P., Sánchez, E., Jiménez, J. I., & Chavarría, M. 2016. Production of
714 selenium nanoparticles in *Pseudomonas putida* KT2440. *Sci. Rep.*, 6(1), 37155.
715 <https://doi.org/10.1038/srep37155>
- 716 Batandjieva, B., Delcheva, T., Duhovnik, B. 2009. Classification of radioactive waste: safety guide: IAEA
717 General Safety Guide GSG-1. Vienna. International Atomic Energy Agency.
- 718 Bengtsson, A., Pedersen, K. 2017. Microbial sulphide-producing activity in water saturated Wyoming MX-
719 80, Asha and Calcigel bentonites at wet densities from 1500 to 2000 kg m⁻³. *Appl. Clay Sci.* 137,
720 203-212. <https://doi.org/10.1016/j.clay.2016.12.024>
- 721 Bienvenu, P., Cassette, P., Andreoletti, G., Bé, M. M., Comte, J., & Lépy, M. C. 2007. A new determination
722 of ⁷⁹Se half-life. *Appl. Radiat. Isot.* 65(3), 355-364. <https://doi.org/10.1016/j.apradiso.2006.09.009>
- 723 Biesinger, M. C. 2017. Advanced analysis of copper X-ray photoelectron spectra. *Surf. Interface Anal.*,
724 49(13), 1325-1334. <https://doi.org/https://doi.org/10.1002/sia.6239>
- 725 Breynaert, E., Scheinost, A.C., Dom, D., Rossberg, A., Vancluysen, J., Gobechiya, E., Kirschhock, C.E.A.,
726 Maes, A. 2010. Reduction of Se(IV) in boom clay: XAS solid phase speciation. *Environ. Sci.*
727 *Technol.*, 44, 6649–6655. <https://doi.org/10.1021/es100569e>
- 728 Chen, S., Wang, P., Zhang, D. 2014. Corrosion behavior of copper under biofilm of sulfate-reducing
729 bacteria. *Corros. Sci.* 87, 407 – 415. <http://dx.doi.org/10.1016/j.corsci.2014.07.001>
- 730 Dou, W., Pu, Y., Han, X., Song, Y., Chen, S., Gu, T. 2020. Corrosion of Cu by a sulfate reducing bacterium
731 in anaerobic vials with different headspace volumes. *Bioelectrochemistry*, 133, 107478.
732 <https://doi.org/10.1016/j.bioelechem.2020.107478>
- 733 El-Batal, A. I., Mosallam, F. M., Ghorab, M. M., Hanora, A., Gobara, M., Baraka, A., ... & El-Sayyad, G.
734 S. 2020. Factorial design-optimized and gamma irradiation-assisted fabrication of selenium
735 nanoparticles by chitosan and *Pleurotus ostreatus* fermented fenugreek for a vigorous in vitro effect

736 against carcinoma cells. *Int J Biol Macromol*, 156, 1584-1599.
737 <https://doi.org/10.1016/j.ijbiomac.2019.11.210>

738 Eswayah, A. S., Smith, T. J., Scheinost, A. C., Hondow, N., & Gardiner, P. H. 2017. Microbial
739 transformations of selenite by methane-oxidizing bacteria. *Appl. Microbiol. Biotechnol.*, 101, 6713-
740 6724. <https://doi.org/10.1007/s00253-017-8380-8>

741 Fairley, N. CasaX.P.S., 2.3.22 ed.; Casa Software Ltd.: 2019.

742 Freikowski, D., Winter, J., & Gallert, C. 2010. Hydrogen formation by an arsenate-reducing *Pseudomonas*
743 *putida*, isolated from arsenic-contaminated groundwater in West Bengal, India. *Appl. Microbiol.*
744 *Biotechnol.*, 88, 1363-1371. <https://doi.org/10.1007/s1201000988520>

745 Garcia-Romero, E., Maríahuert Manchado, E., Suárez, M., and García-Rivas, J. 2019. Spanish bentonites:
746 a review and new data on their geology, mineralogy, and crystal chemistry. *Fortschr. Mineral.* 9:696.
747 <https://doi.org/10.3390/min9110696>

748 Gautam, A., Kushwaha, A., & Rani, R. 2022. Reduction of Hexavalent Chromium [Cr (VI)] by heavy metal
749 tolerant Bacterium *Alkalihalobacillus clausii* CRA1 and its toxicity assessment through flow
750 cytometry. *Curr. Microbiol.*, 79(1), 33. <https://doi.org/10.1007/s00284-021-02734-z>

751 Gebhardt, J. E., McCarron, J. J., Richardson, P. E., & Buckley, A. N. 1986. The effect of cathodic treatment
752 on the anodic polarization of copper sulfides. *Hydrometallurgy*, 17(1), 27-38.
753 [https://doi.org/10.1016/0304-386X\(86\)90018-6](https://doi.org/10.1016/0304-386X(86)90018-6)

754 Ghijsen, J., Tjeng, L. H., van Elp, J., Eskes, H., Westerink, J., Sawatzky, G. A., & Czyzyk, M. T. 1988.
755 Electronic structure of Cu₂O and CuO. *Physical Review B*, 38(16), 11322-11330.
756 <https://doi.org/10.1103/PhysRevB.38.11322>

757 Haghi, M., Diznabi, S. H., Karaboz, I., & Omeroglu, E. E. 2023. Arsenic pollution and arsenic-resistant
758 bacteria of drying Urmia Salt Lake. *Front. Environ. Sci.* 11, 1195643.
759 <https://doi.org/10.3389/fenvs.2023.1195643>

760 Hall, D.S., Behazin, M., Binns, W.J., and Keech, P.G. 2021. An evaluation of corrosion processes affecting
761 copper-coated nuclear waste containers in a deep geological repository. *Prog. Mater. Sci.* 118,
762 100766. <https://doi.org/10.1016/j.pmatsci.2020.100766>

763 Hammer, O., & Harper, D. A. 2001. PAST: paleontological statistics software package for education and
764 data analysis. *Palaeontol. Electron.* <http://palaeo-electronica.org>

765 Hassan, R. S., Abass, M. R., Eid, M. A., & Abdel-Galil, E. A. 2021. Sorption of some radionuclides from
766 liquid waste solutions using anionic clay hydrotalcite sorbent. *Appl. Radiat. Isot.*, 178, 109985.
767 <https://doi.org/10.1016/j.apradiso.2021.109985>

768 He, R., Xia, F. F., Wang, J., Pan, C. L., & Fang, C. R. 2011. Characterization of adsorption removal of
769 hydrogen sulfide by waste biocover soil, an alternative landfill cover. *J. Hazard. Mater.*, 186(1),
770 773-778. <https://doi.org/10.1016/j.jhazmat.2010.11.062>

771 Hoving, A.L., Münch, M.A., Bruggeman, C., Banerjee, D., Behrends, T., 2019. Kinetics of selenite
772 interactions with Boom Clay: adsorption–reduction interplay. *Geol. Soc., Lond., Spec. Publ.* 482,
773 225–239. <https://doi.org/10.1144/SP482-2018-60>.

774 Huertas, F., Fariña, P., Farias, J., García-Siñeriz, J.L., Villar, M.V., Fernández, A.M., Martín, P.L., Elorza,
775 F.J., Gens, A., Sánchez, M., Lloret, A., Samper, J., Martínez, M. Á. 2021. Full-scale Engineered
776 Barriers Experiment. Updated Final Report 1994-2004.

777 Johansson, A.J., Lilja, C., Sjögren, L., Gordon, A., Hallbeck, L., Johansson, L. 2017. Insights from post-
778 test examination of three packages from the MiniCan test series of copper cast iron canisters for
779 geological disposal of spent nuclear fuel: impact of the presence and density of bentonite clay.
780 *Corros. Eng. Sci. Technol.* 52, 54-60. <https://doi.org/10.1080/1478422X.2017.1296224>

781 Jolley, J. G., Geesey, G. G., Hankins, M. R., Wright, R. B., & Wichlacz, P. L. 1989. Auger electron and X-
782 ray photoelectron spectroscopic study of the biocorrosion of copper by alginic acid polysaccharide.
783 *Appl. Surf. Sci.*, 37(4), 469-480. [https://doi.org/https://doi.org/10.1016/0169-4332\(89\)90505-9](https://doi.org/https://doi.org/10.1016/0169-4332(89)90505-9)

784 Jörg, G., Bühnemann, R., Hollas, S., Kivel, N., Kossert, K., Van Winckel, S., & Gostomski, C. L. V. 2010.
785 Preparation of radiochemically pure ⁷⁹Se and highly precise determination of its half-life. *Appl.*
786 *Radiat. Isot.*, 68(12), 2339-2351. <https://doi.org/10.1016/j.apradiso.2010.05.006>

787 Kang, M., Chen, F., Wu, S., Yang, Y., Bruggeman, C., Charlet, L. 2011. Effect of pH on Aqueous Se(IV)
788 Reduction by Pyrite. *Environ. Sci. Technol.* 45, 2704–2710. <https://doi.org/10.1021/es1033553>.

789 Keech, P. G., Vo, P., Ramamurthy, S., Chen, J., Jacklin, R., Shoesmith, D. W. 2014. Design and development
790 of copper coatings for long term storage of used nuclear fuel. *Corros. Eng. Sci. Technol.*, 49(6), 425-
791 430. <https://doi.org/10.1179/1743278214Y.0000000206>

792 Le Caër, S. 2011. Water radiolysis: influence of oxide surfaces on H₂ production under ionizing radiation.
793 *Water*, 3(1), 235-253. <https://doi.org/10.3390/w3010235>

794 Liu, D., Dong, H., Bishop, M.E., Zhang, J., Wang, H., Xie, S., Wang, S., Huang, L., Eberl, D.D. 2012.
795 Microbial reduction of structural iron in interstratified illite-smectite minerals by a sulfate-reducing
796 bacterium. *Geobiology*, 10, 150-162. <https://doi.org/10.1111/j.1472-4669.2011.00307.x>

797 Mansour, A. N. 1994. Copper Mg K α XPS Spectra from the Physical Electronics Model 5400 Spectrometer.
798 *Surf. Sci. Spectra*, 3(3), 202-210. <https://doi.org/10.1116/1.1247748>

799 Martinez-Moreno, M. F., Povedano-Priego, C., Mumford, A. D., Morales-Hidalgo, M., Mijndonckx, K.,
800 Jroundi, F., Ojeda, J. J., & Merroun, M. L. 2024a. Microbial responses to elevated temperature:
801 Evaluating bentonite mineralogy and copper canister corrosion within the long-term stability of deep
802 geological repositories of nuclear waste. *Sci. Total Environ.*, 170149.
803 <https://doi.org/10.1016/j.scitotenv.2024.170149>

804 Martinez-Moreno, M. F., Povedano-Priego, C., Morales-Hidalgo, M., Mumford, A. D., Aranda, E., Vilchez-
805 Vargas, R., ... & Merroun, M. L. 2024b. Microbial influence in Spanish bentonite slurry microcosms:
806 unveiling a-year long geochemical evolution and early-stage copper corrosion related to nuclear
807 waste repositories. *Env. Poll.*, 124491. <https://doi.org/10.1016/j.envpol.2024.124491>

808 Martinez-Moreno, M.F., Povedano-Priego, C., Morales-Hidalgo, M., Mumford, A.D., Ojeda, J.J., Jroundi,
809 F., Merroun, M. L. 2023. Impact of compacted bentonite microbial community on the clay
810 mineralogy and copper canister corrosion: a multidisciplinary approach in view of a safe Deep
811 Geological Repository of nuclear wastes. *J. Hazard. Mater.* 131940.
812 <https://doi.org/10.1016/j.jhazmat.2023.131940>

813 Masurat, P., Eriksson, S., Pedersen, K. 2010. Microbial sulphide production in compacted Wyoming
814 bentonite MX-80 under in situ conditions relevant to a repository for high-level radioactive
815 waste. *Appl. Clay Sci.* 47, 58-64. <https://doi.org/10.1016/j.clay.2009.01.004>

816 Matschiavelli, N., Kluge, S., Podlech, C., Standhaft, D., Grathoff, G., Ikeda-Ohno, A., ... & Cherkouk, A.
817 2019. The year-long development of microorganisms in uncompacted bavarian bentonite slurries at
818 30 and 60 °C. *Environ. Sci. Technol.* 2019, 53, 10514–10524.
819 <https://doi.org/10.1021/acs.est.9b02670>

820 Meleshyn, A. 2011. Microbial processes relevant for long-term performance of radioactive waste
821 repositories in clays. GRS-291. ISBN 978-3-939355-67-0. Available from:
822 <https://www.grs.de/sites/default/files/pdf/GRS-291.pdf>

823 Miettinen, H., Bomberg, M., Bes, R., Tiljander, M., & Vikman, M. 2022. Transformation of inherent
824 microorganisms in Wyoming-type bentonite and their effects on structural iron. *Appl. Clay Sci.*, 221,
825 106465. <https://doi.org/10.1016/j.clay.2022.106465>

826 Morales-Hidalgo, M., Povedano-Priego, C., Martinez-Moreno, M. F., Ruiz-Fresneda, M. A., Lopez-
827 Fernandez, M., Jroundi, F., & Merroun, M. L. 2024. Insights into the Impact of Physicochemical
828 and Microbiological Parameters on the Safety Performance of Deep Geological Repositories.
829 *Microorganisms*, 12(5), 1025. <https://doi.org/10.3390/microorganisms12051025>

830 Moulder, J.F., Stickle, W.F., Sobol, P.E., Bomben, K.D. 1992. Handbook of X-Ray Photoelectron
831 Spectroscopy, Perkin-Elmer Corporation, Physical Electronics Division, Eden Prairie, Minn.

832 Park, S. Y., Zhang, Y., O’Loughlin, E. J., Jo, H. Y., Kwon, J. S., & Kwon, M. J. 2024. Temperature-
833 dependent microbial reactions by indigenous microbes in bentonite under Fe (III)-and sulfate-
834 reducing conditions. *J. Hazard. Mater.*, 465, 133318. <https://doi.org/10.1016/j.jhazmat.2023.133318>

835 Patel, S., & Gupta, R. S. 2020. A phylogenomic and comparative genomic framework for resolving the
836 polyphyly of the genus *Bacillus*: Proposal for six new genera of *Bacillus* species, *Peribacillus* gen.
837 nov., *Cytobacillus* gen. nov., *Mesobacillus* gen. nov., *Neobacillus* gen. nov., *Metabacillus* gen. nov.
838 and *Alkalihalobacillus* gen. nov. *Int. J. Syst. Evol. Microbiol.*, 70(1), 406-438.
839 <https://doi.org/10.1099/ijsem.0.003775>

840 Payer, J. H., Finsterle, S., Apps, J.A., Muller, R.A. 2019. Corrosion performance of engineered barrier
841 system in deep horizontal drillholes. *Energies*. 12, 1491. <https://doi.org/10.3390/en12081491>

842 Pentráková, L., Su, K., Pentrák, M., Stucki, J.W. 2013. A review of microbial redox interactions with
843 structural Fe in clay minerals. *Clay Miner.* 48, 543-560.
844 <https://doi.org/10.1180/claymin.2013.048.3.10>

845 Povedano-Priego, C., Jroundi, F., Solari, P. L., Guerra-Tschuschke, I., del Mar Abad-Ortega, M., Link, A.,
846 ... & Merroun, M. L. 2023. Unlocking the bentonite microbial diversity and its implications in
847 selenium bioreduction and biotransformation: Advances in deep geological repositories. *J. Hazard.*
848 *Mater.*, 445, 130557. <https://doi.org/10.1016/j.jhazmat.2022.130557>

849 Povedano-Priego, C., Jroundi, F., Lopez-Fernandez, M., Shrestha, R., Spanek, R., Martín-Sánchez, I.,
850 Villar, M.V., Ševců, A., Dopson, M. Merroun, M.L. 2021. Deciphering indigenous bacteria in
851 compacted bentonite through a novel and efficient DNA extraction method: Insights into

852 biogeochemical processes within the Deep Geological Disposal of nuclear waste concept. *J. Hazard.*
853 *Mater.* 408, 124600. <https://doi.org/10.1016/j.jhazmat.2020.124600>

854 R Core Team. 2022. R: A language and environment for statistical computing. R Foundation for Statistical
855 Computing, Vienna, Austria. URL <https://www.Rproject.org/>

856 Robertson, C.E., Harris, J.K., Wagner, B.D., Granger, D., Browne, K., Tatem, B., Feazel, L.M., Park, K.,
857 Pace, N.R., Frank, D.N. 2013. Explicet: graphical user interface software for metadata-driven
858 management, analysis and visualization of microbiome data. *Bioinformatics*, 29, 3100-3101.
859 <https://doi.org/10.1093/bioinformatics/btt526>

860 Romand, M., Roubin, M., & Deloume, J. P. 1978. ESCA studies of some copper and silver selenides. J.
861 *Electron Spectros. Relat. Phenomena*, 13(3), 229-242. [https://doi.org/10.1016/0368-](https://doi.org/10.1016/0368-2048(78)85029-4)
862 [2048\(78\)85029-4](https://doi.org/10.1016/0368-2048(78)85029-4)

863 Ruiz-Fresneda, M. A., Martinez-Moreno, M. F., Povedano-Priego, C., Morales-Hidalgo, M., Jroundi, F., &
864 Merroun, M. L. 2023. Impact of microbial processes on the safety of deep geological repositories
865 for radioactive waste. *Front. Microbiol.*, 14, 1134078. <https://doi.org/10.3389/fmicb.2023.1134078>

866 Ruiz-Fresneda, M. A., Staicu, L. C., Lazuén-López, G., & Merroun, M. L. 2023b. Allotropy of selenium
867 nanoparticles: Colourful transition, synthesis, and biotechnological applications. *Microb.*
868 *Biotechnol.*, 16(5), 877-892. <https://doi.org/10.1111/1751-7915.14209>

869 Ruiz-Fresneda, M. A., Eswayah, A. S., Romero-González, M., Gardiner, P. H., Solari, P. L., & Merroun, M.
870 L. 2020. Chemical and structural characterization of Se^{IV} biotransformations by *Stenotrophomonas*
871 *bentonitica* into Se⁰ nanostructures and volatile Se species. *Environ. Sci. Nano*, 7(7), 2140-2155.
872 <https://doi.org/10.1039/D0EN00507J>

873 Ruiz-Fresneda, M. A., Gomez-Bolivar, J., Delgado-Martin, J., Abad-Ortega, M. D. M., Guerra-Tschuschke,
874 I., & Merroun, M. L. 2019. The bioreduction of selenite under anaerobic and alkaline conditions
875 analogous to those expected for a deep geological repository system. *Molecules*, 24(21), 3868.
876 <https://doi.org/10.3390/molecules24213868>

877 Ruiz-Fresneda, M. A., Delgado-Martín, J., Gomez-Bolívar, J., Fernandez-Cantos, M. V., Bosch-Estévez,
878 G., Martinez-Moreno, M. F., & Merroun, M. L. 2018. Green synthesis and biotransformation of
879 amorphous Se nanospheres to trigonal 1D Se nanostructures: impact on Se mobility within the
880 concept of radioactive waste disposal. *Environ. Sci. Nano*, 5(9), 2103-2116.
881 <https://doi.org/10.1039/C8EN00221E>

- 882 Salehi Alaei, E., Guo, M., Chen, J., Behazin, M., Bergendal, E., Lilja, C., ... & Noël, J. J. 2023. The
883 transition from used fuel container corrosion under oxic conditions to corrosion in an anoxic
884 environment. *Mater. Corros.*, 74(11-12), 1690-1706. <https://doi.org/10.1002/maco.202313757>
- 885 Sanchez-Castro, I., Ruiz-Fresneda, M. A., Bakkali, M., Kämpfer, P., Glaeser, S. P., Busse, H. J., ... &
886 Merroun, M. L. 2017. *Stenotrophomonas bentonitica* sp. nov., isolated from bentonite formations.
887 *Int. J. Syst. Evol. Microbiol.*, 67(8), 2779. <https://doi.org/10.1099%2Fijs.0.002016>
- 888 Shelobolina, E.S., VanPraagh, C.G., Lovley, D.R. 2003. Use of ferric and ferrous iron containing minerals
889 for respiration by *Desulfotobacterium frappieri*. *Geomicrobiol. J.* 20, 143-156.
890 <https://doi.org/10.1080/01490450303884>
- 891 Shiratori-Takano, H., Akita, K., Yamada, K., Itoh, T., Sugihara, T., Beppu, T., & Ueda, K. 2014. Description
892 of *Symbiobacterium ostreiconchae* sp. nov., *Symbiobacterium turbinis* sp. nov. and
893 *Symbiobacterium terraclitae* sp. nov., isolated from shellfish, emended description of the genus
894 *Symbiobacterium* and proposal of *Symbiobacteriaceae* fam. nov. *Int. J. Syst. Evol. Microbiol.*,
895 64(Pt_10), 3375-3383. <https://doi.org/10.1099/ijms.0.063750-0>
- 896 Sitaud, B., Solari, P. L., Schlutig, S., Llorens, I., & Hermange, H. 2012. Characterization of radioactive
897 materials using the MARS beamline at the synchrotron SOLEIL. *J. Nucl. Mater.*, 425(1-3), 238-
898 243. <https://doi.org/10.1016/j.jnucmat.2011.08.017>
- 899 Stackebrandt, E., Sproer, C., Rainey, F. A., Burghardt, J., Päufer, O., & Hippe, H. 1997. Phylogenetic
900 analysis of the genus *Desulfotomaculum*: evidence for the misclassification of *Desulfotomaculum*
901 *guttoideum* and description of *Desulfotomaculum orientis* as *Desulfosporosinus orientis* gen. nov.,
902 comb. nov. *Int. J. Syst. Evol. Microbiol.*, 47(4), 1134-1139. [https://doi.org/10.1099/00207713-47-4-
903 1134](https://doi.org/10.1099/00207713-47-4-1134)
- 904 Staicu, L. C., & Barton, L. L. 2021. Selenium respiration in anaerobic bacteria: Does energy generation pay
905 off?. *J. Inorg. Biochem.*, 222, 111509. <https://doi.org/10.1016/j.jinorgbio.2021.111509>
- 906 Staicu, L. C., Ackerson, C. J., Cornelis, P., Ye, L., Berendsen, R. L., Hunter, W. J., ... & Pilon-Smits, E. A.
907 2015. *Pseudomonas moraviensis* subsp. *stanleyae*, a bacterial endophyte of hyperaccumulator
908 *Stanleya pinnata*, is capable of efficient selenite reduction to elemental selenium under aerobic
909 conditions. *J. Appl. Microbiol.*, 119(2), 400-410. <https://doi.org/10.1111/jam.12842>
- 910 Villar, M.V., Fernández-Soler, J.M., Delgado Huertas, A., Reyes, E., Linares, J., Jiménez de Cisneros, C.,
911 Linares, J., Reyes, E., Delgado, A., Fernandez-Soler, J.M., Astudillo, J. 2006. The study of Spanish

912 clays for their use as sealing materials in nuclear waste repositories: 20 years of progress. *J. Iber.*
913 *Geol.* 32, 15-36.

914 Vogel, M., Fischer, S., Maffert, A., Hübner, R., Scheinost, A. C., Franzen, C., & Steudtner, R. 2018.
915 Biotransformation and detoxification of selenite by microbial biogenesis of selenium-sulfur
916 nanoparticles. *J. Hazard. Mater.*, 344, 749-757. <https://doi.org/10.1016/j.jhazmat.2017.10.034>

917 Wagner, C.D., Riggs, W.M., Davis, L.E., Moulder, J.F., Muilenberg, G.E. 1979. Handbook of x-ray
918 photoelectron spectroscopy: a reference book of standard data for use in x-ray photoelectron
919 spectroscopy. Perkin-Elmer Corporation: Eden Prairie, MN.

920 Wang, D., Rensing, C., & Zheng, S. 2022. Microbial reduction and resistance to selenium: Mechanisms,
921 applications and prospects. *J. Hazard. Mater.*, 421, 126684.
922 <https://doi.org/10.1016/j.jhazmat.2021.126684>

923 Weser, U., Sokolowski, G., & Pilz, W. 1977. Reaction of selenite with biochemically active thiols: An X-
924 ray photoelectron spectroscopic study. *J. Electron Spectros. Relat. Phenomena*, 10(4), 429-439.
925 [https://doi.org/10.1016/0368-2048\(77\)85039-1](https://doi.org/10.1016/0368-2048(77)85039-1)

926 WNA. 2024. World Nuclear Association. Storage and disposal of radioactive waste. [https://www.world-](https://www.world-nuclear.org/)
927 [nuclear.org/](https://www.world-nuclear.org/)

928

Table 1.

Sample ID and content of the different microcosms. All microcosms contained each three Cu mini-canisters (Cu-mCan) and 2 mM of Se(IV). The microcosms were set up in triplicate and anoxically incubated at 28 °C. ALS: acetate, lactate and sulfate concentrations; BPAS: bacterial consortium; +: presence; -: absence.

Sample ID	Bentonite	ALS (mM)	BPAS	Se(IV)
B.eD.Se.BC	Non-heat-shocked	30:10:20	+	2 mM
B.Se.BC		-	+	
B.eD.Se		30:10:20	-	
B.Se		-	-	
StB.eD.Se.BC	Heat-shocked	30:10:20	+	
StB.Se.BC		-	+	
StB.eD.Se		30:10:20	-	
StB.Se		-	-	

Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

Table 2. EXAFS structural parameters of the Se foil and the selenium reduction products (SeRPs) from the selected Se-amended microcosms after 45 days of incubation.

Sample	Shell	N^a	R^b [Å]	σ^{2c} [Å ²]	ΔE [eV]
Se foil	Se–Se	3.2 ± 0.2	2.37	0.0040	–10.10
B.eD.Se.BC	Se–Se	1.9 ± 0.2	2.33	0.0013	–13.70
B.Se.BC	Se–Se	2.7 ± 0.2	2.34	0.0026	–12.37
StB.eD.Se.BC	Se–Se	2.4 ± 0.1	2.35	0.0024	–11.10
StB.Se.BC	Se–Se	2.5 ± 0.1	2.35	0.0025	–11.60

^aErrors in coordination numbers are $\pm 25\%$ and standard deviations as estimated by EXAFSPAK. ^bErrors in distance are ± 0.02 Å. ^cDebye–Waller factor. Glossary: B: bentonite, StB: heat-shocked bentonite, eD: addition of electron donors/acceptor (ALS), Se: 2 mM of Se(IV), BC: spiked with BPAS consortium.

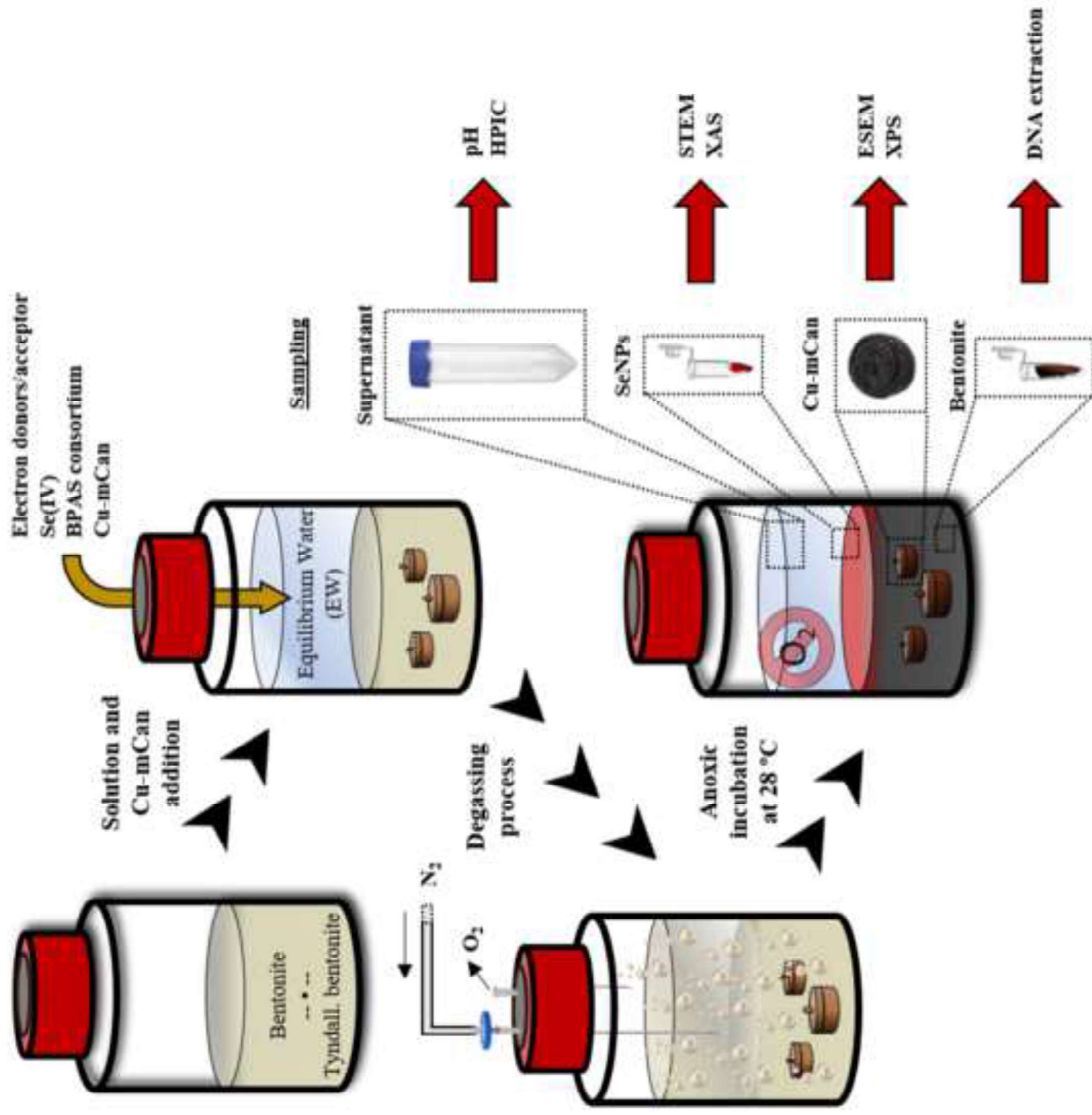


Figure 2

[Click here to access/download;Figure;Figure 2.TIF](#)

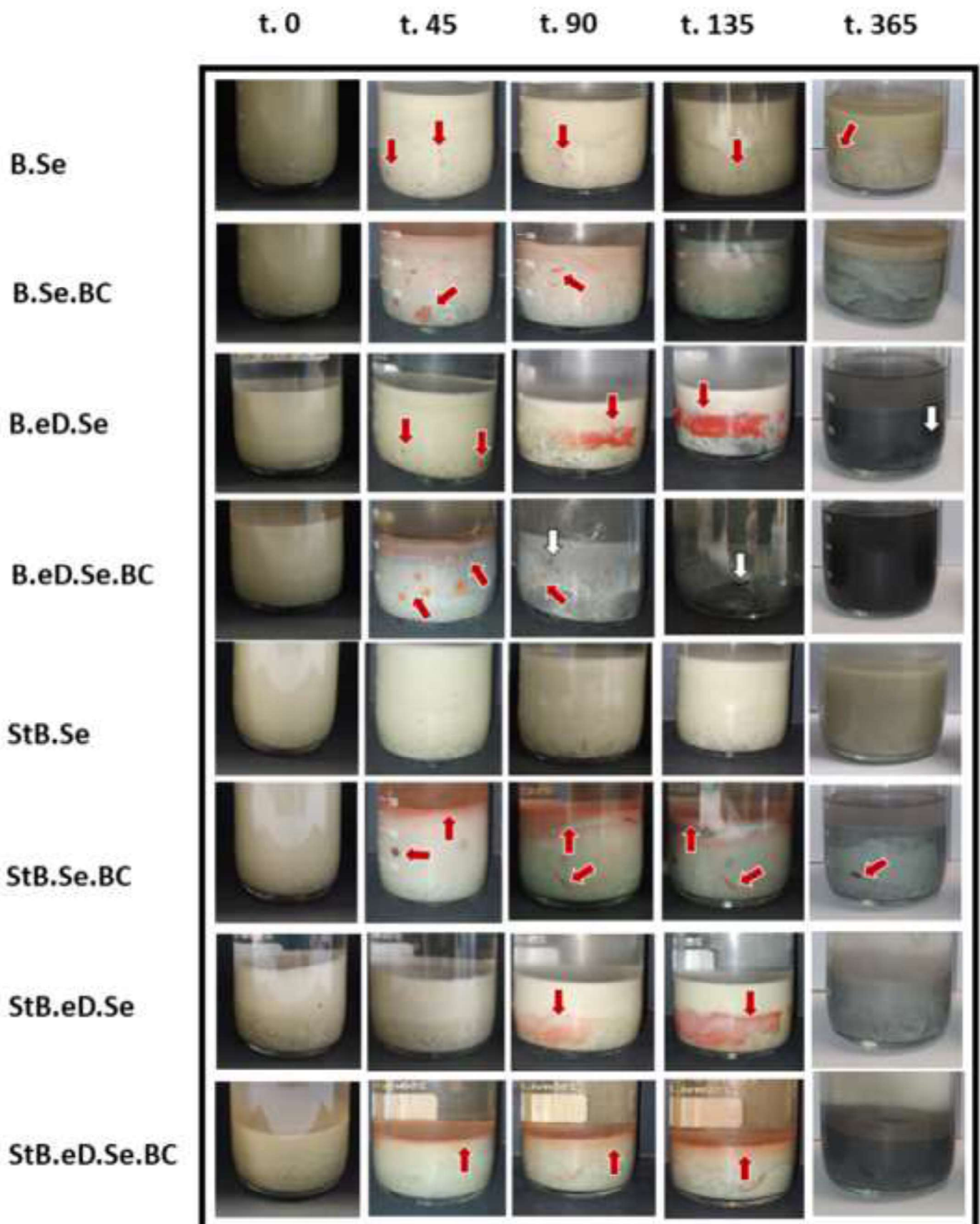
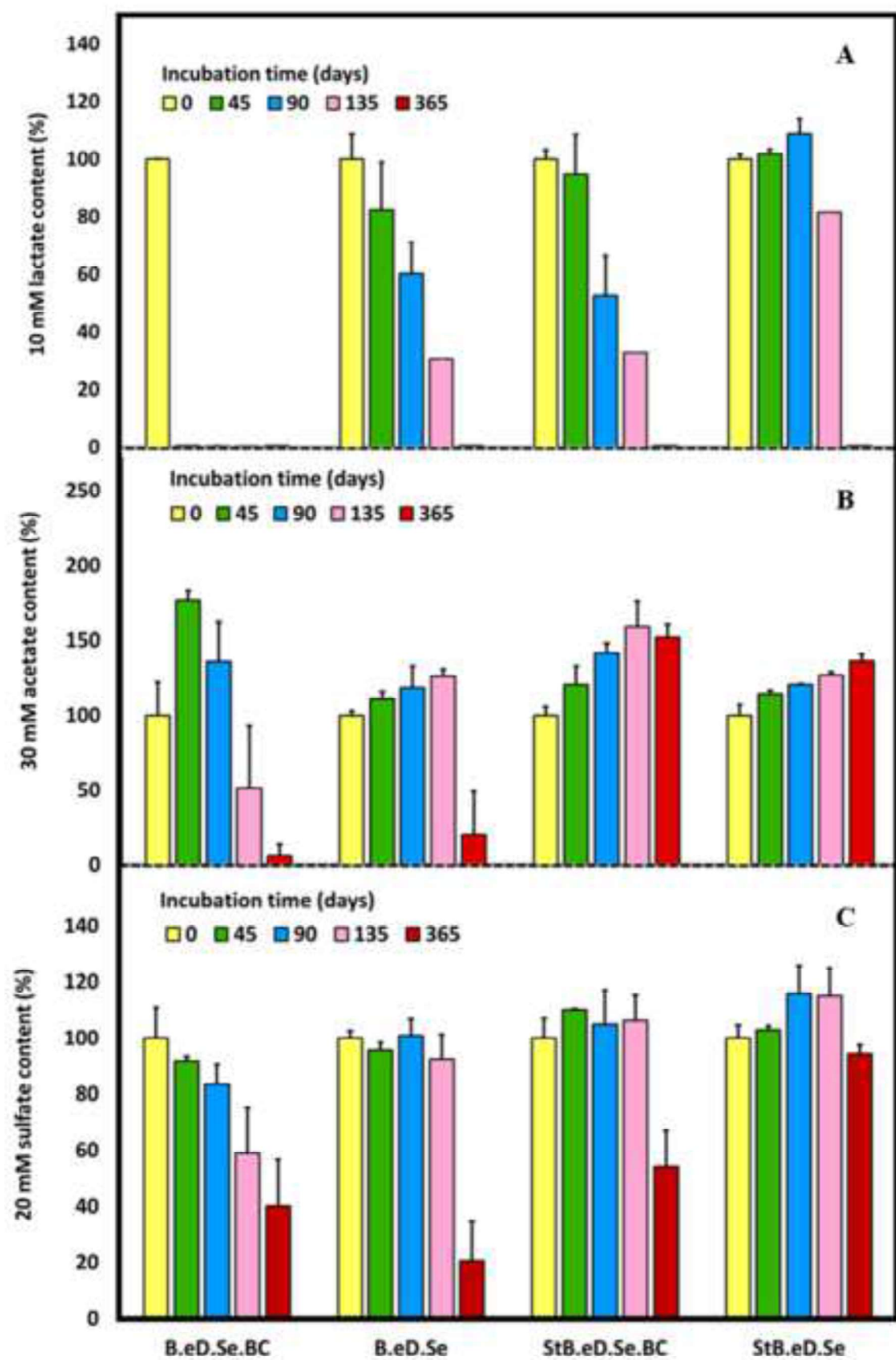
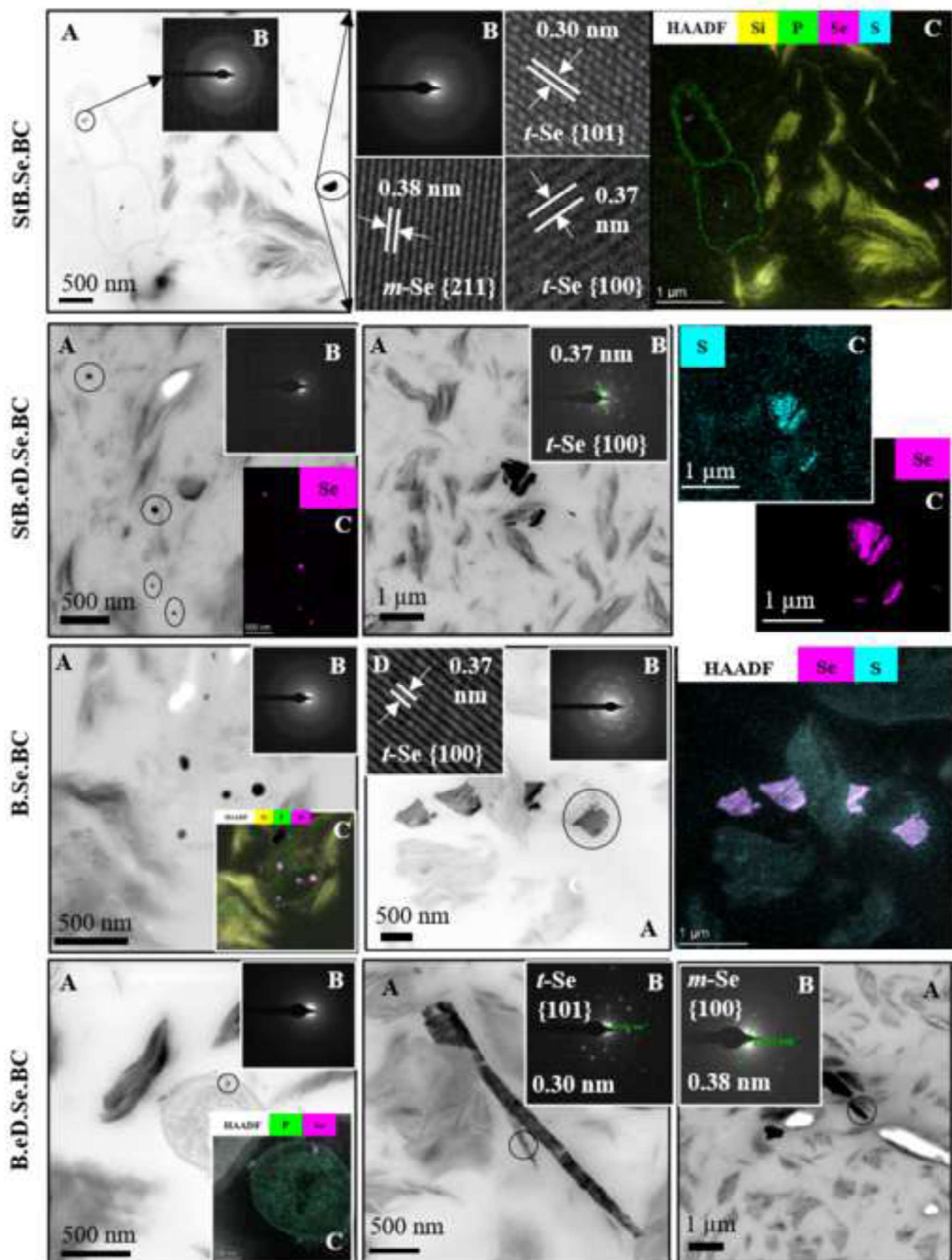
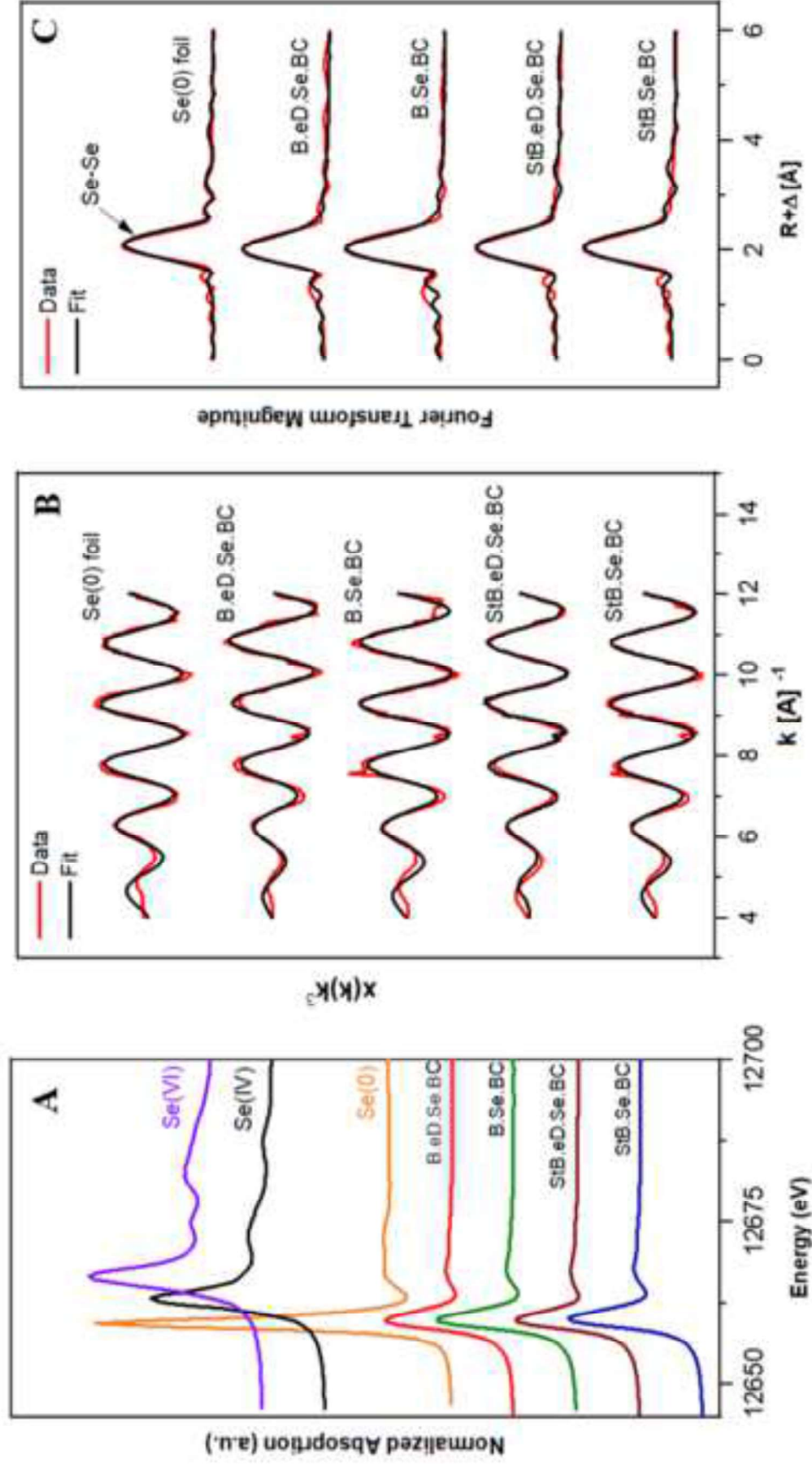


Figure 3

[Click here to access/download;Figure;Figure 3.TIF](#)







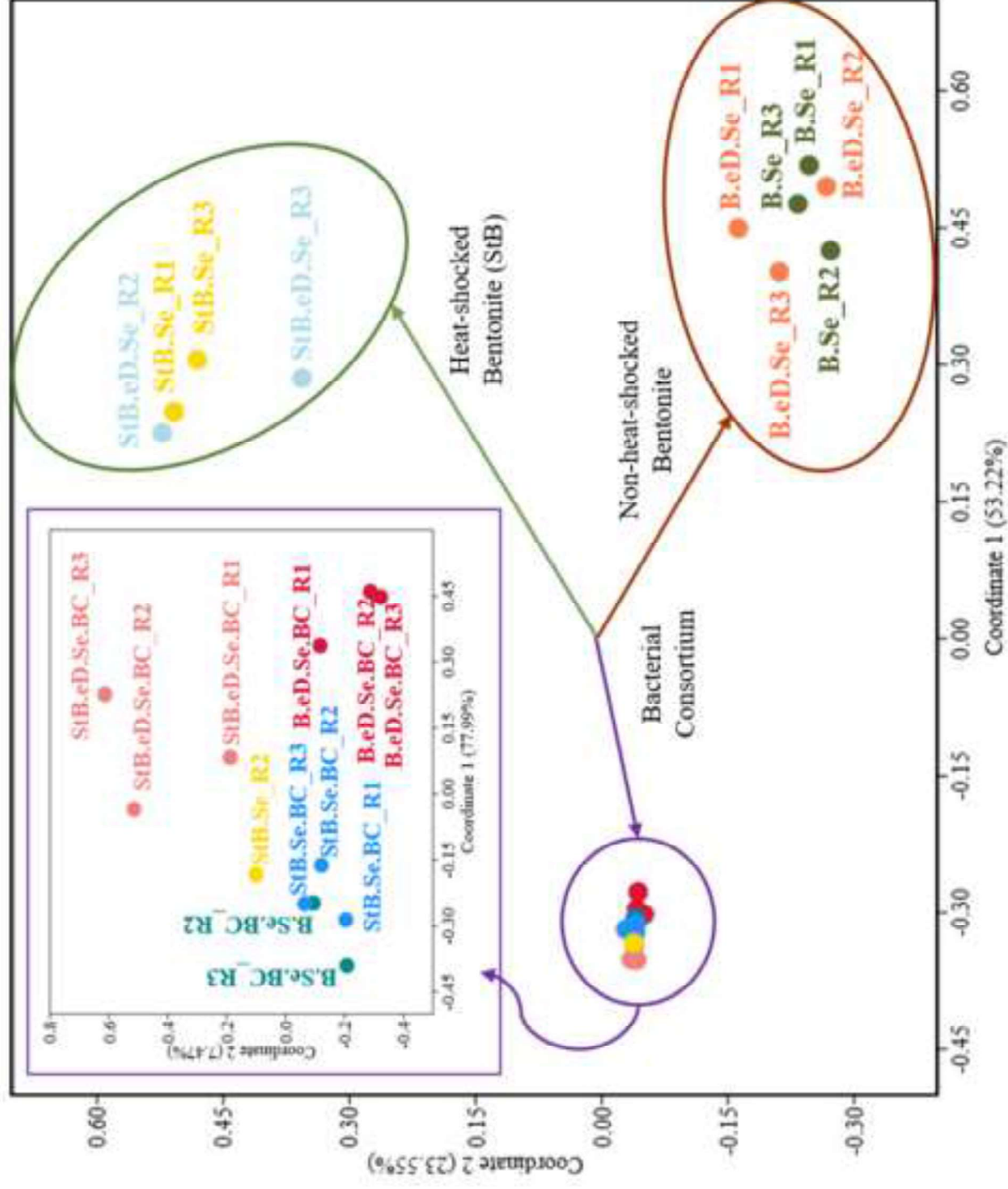
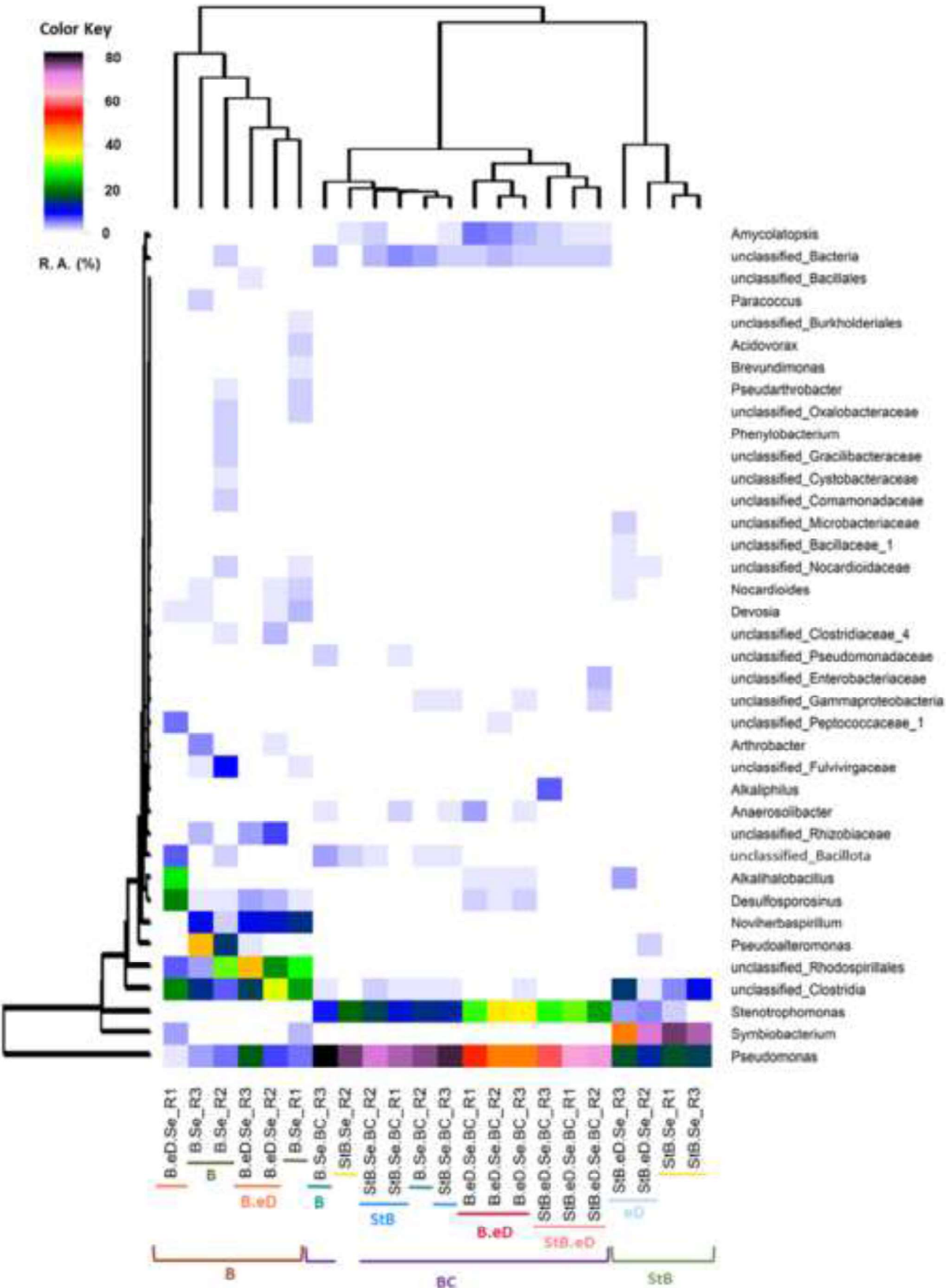


Figure 7

[Click here to access/download;Figure;Figure 7.TIF](#)



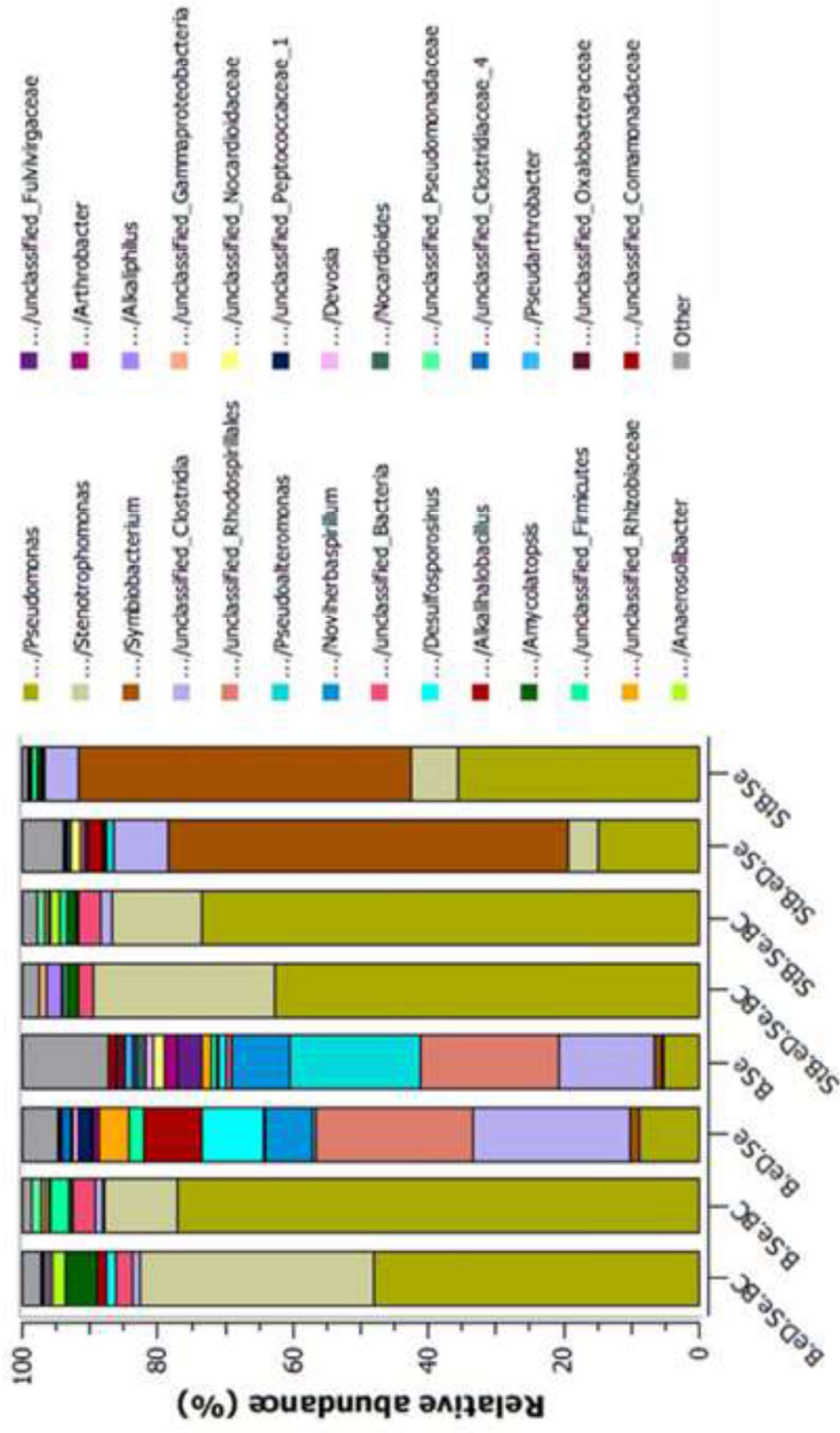


Figure 8

